Amendment to the

Upper Delaware Water Quality Management Plan Northeast Water Quality Management Plan Upper Raritan Water Quality Management Plan Sussex County Water Quality Management Plan

Total Maximum Daily Loads for Phosphorus To Address 4 Eutrophic Lakes in the Northwest Water Region

CRANBERRY LAKE, SUSSEX COUNTY
GHOST LAKE, WARREN COUNTY
LAKE HOPATCONG, SUSSEX COUNTY
LAKE MUSCONETCONG, SUSSEX COUNTY

Watershed Management Area 1 (Upper Delaware River Watershed)

Proposed: January 21, 2003

Established: March 28, 2003

Approved (by EPA Region 2): September 17, 2003

Adopted:

New Jersey Department of Environmental Protection Division of Watershed Management P.O. Box 418 Trenton, New Jersey 08625-0418

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1.0 Executive Summary

The State of New Jersey's 2002 Integrated List of Waterbodies identified several lakes in the Northwest Water Region as being eutrophic. This report establishes total maximum daily loads (TMDLs) for total phosphorus (TP) that address eutrophication of the lakes listed in Table 1.

Table 1 Eutrophic Lakes for which Phosphorus TMDLs are being established

| TMDL Number | Lake Name | Municipality | WMA | Acres |
|----------------|-------------------|---|-----|-------|
| 1 | Cranberry Lake | Byram Township, Sussex County | 01 | 190 |
| 2 | Ghost Lake | Independence Township, Warren County | 01 | 18.3 |
| 3 | Lake Hopatcong | Hopatcong Borough, Sussex County; Mt. Arlington Borough, Jefferson & Roxbury Townships; Morris County | 01 | 2,410 |
| 4 | Lake Musconetcong | Stanhope, Byram, Netcong and Roxbury Townships; Sussex County | 01 | 314 |

These TMDLs serve as the foundation on which restoration plans will be developed to restore eutrophic lakes and thereby attain applicable surface water quality standards. A TMDL is developed as a mechanism for identifying all the contributors to surface water quality impacts and setting goals for load reductions for pollutants of concern as necessary to meet Surface Water Quality Standards (SWQS). The pollutant of concern for these TMDLs is phosphorus, since phosphorus is generally the nutrient responsible for overfertilization of inland lakes leading to cultural eutrophication. The Department's Geographic Information System (GIS) was used extensively to describe the lakes and lakesheds (drainage basins of the lakes).

In order to prevent excessive primary productivity¹ and consequent impairment of recreational, water supply and aquatic life designated uses, the SWQS define both numerical and narrative criteria that address eutrophication in lakes due to overfertilization. Phosphorus sources were characterized on an annual scale (kg TP/yr) for both point and nonpoint sources. Runoff from land surfaces comprises a substantial source of phosphorus into lakes. An empirical model was used to relate annual phosphorus load and steady-state in-lake concentration of total phosphorus. To achieve the TMDLs, overall load reductions were calculated for at least eight source categories. In order to track effectiveness of remediation measures (including TMDLs) and to develop baseline and trend information on lakes, the Department will augment its ambient monitoring program to include lakes on a rotating schedule. The implementation plan also calls for the collection of additional monitoring data and the development of a Lake Restoration Plan for each lake for which TMDLs are being established. These plans will consider what specific measures are necessary to achieve the nutrient reductions required by the TMDL, as well as what in-lake measures need to be taken to supplement the nutrient reductions required by the TMDL.

of the food web.

5

Primary productivity refers to the growth rate of primary producers, namely algae and aquatic plants, which form the base

Each TMDL shall be proposed and adopted by the Department as an amendment to the appropriate areawide water quality management plan(s) in accordance with N.J.A.C. 7:15-3.4(g).

This TMDL Report is consistent with EPA's May 20, 2002 guidance document entitled: "Guidelines for Reviewing TMDLs under Existing Regulations issued in 1992," (Suftin, 2002) which describes the statutory and regulatory requirements for approvable TMDLs.

2.0 Introduction

Sublist 5 (also known as List 5 or, traditionally, the 303(d) List) of the State of New Jersey's 2002 Integrated List of Waterbodies identified several lakes in the Northwest Water Region (WMAs 1, 2, and 11) as being eutrophic, as evidenced by elevated total phosphorus (TP), elevated chlorophyll-a, and/or macrophyte density that impairs recreational use (a qualitative assessment). Total phosphorus was used as the pollutant of concern, since this "independent" causal pollutant causes "dependent " responses in chlorophyll-a concentrations and/or macrophyte density. This report establishes four total maximum daily loads (TMDLs) that address total phosphorus loads to the identified lakes. These TMDLs serve as the foundation on which management approaches or restoration plans will be developed to restore eutrophic lakes and thereby attain applicable surface water quality standards. Several of the lakes are listed on Sublist 5 for impairments caused by other pollutants. These TMDLs address only the impairment of lakes due to eutrophication. Separate TMDL evaluations will be developed to address the other pollutants of concern. The waterbodies will remain on Sublist 5 until such time as TMDL evaluations for all pollutants have been completed and approved by the United States Environmental Protection Agency (USEPA).

A TMDL is considered "proposed" when NJDEP publishes the TMDL Report as a proposed Water Quality Management Plan Amendment in the New Jersey Register (NJR) for public review and comment. A TMDL is considered to be "established" when NJDEP finalizes the TMDL Report after considering comments received during the public comment period for the proposed plan amendment and formally submits it to EPA Region 2 for thirty (30)-day review and approval. The TMDL is considered "approved" when the NJDEP-established TMDL is approved by EPA Region 2. The TMDL is considered to be "adopted" when the EPA-approved TMDL is adopted by NJDEP as a water quality management plan amendment and the adoption notice is published in the NJR.

3.0 Background

3.1 305(b) Report and 303(d) List

In accordance with Section 305(b) of the Federal Clean Water Act (CWA) (33 U.S.C. 1315(B)), the State of New Jersey is required to biennially prepare and submit to the United States

Environmental Protection Agency (USEPA) a report addressing the overall water quality of the State's waters. This report is commonly referred to as the 305(b) Report or the Water Quality Inventory Report.

In accordance with Section 303(d) of the CWA, the State is also required to biennially prepare and submit to USEPA a report that identifies waters that do not meet or are not expected to meet surface water quality standards (SWQS) after implementation of technology-based effluent limitations or other required controls. This report is commonly referred to as the 303(d) List. The listed waterbodies are considered water quality-limited and require total maximum daily load (TMDLs) evaluations. For waterbodies identified on the 303(d) List, there are three possible scenarios that may result in a waterbody being removed from the 303(d) List:

Scenario 1: A TMDL is established for the pollutant of concern;

Scenario 2: A determination is made that the waterbody is meeting water quality standards (no TMDL is required); or

Scenario 3: A determination is made that a TMDL is not the appropriate mechanism for achieving water quality standards and that other control actions will result in meeting standards.

Where a TMDL is required (Scenario 1), it will: 1) specify the maximum amount of a pollutant that a waterbody can receive and still meet water quality standards; and 2) allocate pollutant loadings among point and nonpoint pollutant sources.

Recent EPA guidance (Suftin, 2002) describes the statutory and regulatory requirements for approvable TMDLs, as well as additional information generally needed for USEPA to determine if a submitted TMDL fulfills the legal requirements for approval under Section 303(d) and EPA regulations. The Department believes that this TMDL report, which includes four TMDLs, addresses the following items in the May 20, 2002 guideline document:

- 1. Identification of waterbody(ies), pollutant of concern, pollutant sources and priority ranking.
- 2. Description of applicable water quality standards and numeric water quality target(s).
- 3. Loading capacity linking water quality and pollutant sources.
- 4. Load allocations.
- 5. Wasteload allocations.
- 6. Margin of safety.
- 7. Seasonal variation.
- 8. Reasonable assurances.
- 9. Monitoring plan to track TMDL effectiveness.
- 10. Implementation (USEPA is not required to and does not approve TMDL implementation plans).
- 11. Public Participation.
- 12. Submittal letter.

3.2 Total Maximum Daily Loads (TMDLs)

A TMDL represents the assimilative or carrying capacity of a waterbody, taking into consideration point and nonpoint source of pollutants of concern, natural background and surface water withdrawals. A TMDL quantifies the amount of a pollutant a water body can assimilate without violating a state's water quality standards and allocates that load capacity to known point sources in the form of wasteload allocations (WLAs), nonpoint sources in the form of load allocations (LAs), and a margin of safety. A TMDL is developed as a mechanism for identifying all the contributors to surface water quality impacts and setting goals for load reductions for pollutants of concern as necessary to meet SWQS.

Once one of the three possible delisting scenarios, noted above, is completed, states have the option to remove the waterbody and specific pollutant of concern from the 303(d) List or maintain the waterbody on the 303(d) list until SWQS are achieved. The State of New Jersey will be removing lakes from the 303(d) List for eutrophication once their TMDLS are approved by USEPA.

3.3 Integrated List of Waterbodies

In November 2001, USEPA issued guidance that encouraged states to integrate the 305(b) Report and the 303(d) List into one report. This integrated report assigns waterbodies to one of five categories. In general, Categories 1 through 4 include a range of designated use impairments with a discussion of enforceable management strategies, whereas Sublist 5 constitutes the traditional 303(d) List for waters impaired or threatened by a pollutant for which one or more TMDL evaluations are needed. Where more than one pollutant is associated with the impairment for a given waterbody, that waterbody will remain on Sublist 5 until one of the three possible delisting scenarios is completed. In the case of an Integrated List, however, the waterbody is not delisted but moved to one of the other categories.

Following USEPA's guidance, the Department chose to develop an Integrated Report for New Jersey. New Jersey's 2002 Integrated List of Waterbodies is based upon these five categories and identifies water quality limited surface waters in accordance with N.J.A.C. 7:15-6 and Section 303(d) of the CWA. These TMDLs address eutrophic lakes, as listed on Sublist 5 of the State of New Jersey's 2002 Integrated List of Waterbodies.

4.0 Pollutant of Concern and Area of Interest

Lakes were designated as eutrophic on Sublist 5 of the 2002 Integrated List of Waterbodies as a result of evaluations performed through the State's Clean Lakes Program. Indicators used to determine trophic status included elevated total phosphorus (TP), elevated chlorophyll-a, and/or macrophyte density. The pollutant of concern for these TMDLs is total phosphorus. The mechanism by which phosphorus can cause use impairment is via excessive primary productivity. Phosphorus is an essential nutrient for plants and algae, but is considered a pollutant because it can stimulate excessive growth (primary production). Phosphorus is

most often the major nutrient in shortest supply relative to the nutritional requirements of primary producers in freshwater lakes; consequently, phosphorus is frequently a prime determinant of the total biomass in a lake. Furthermore, of the major nutrients, phosphorus is the most effectively controlled through engineering technology and land use management (Holdren *et al*, 2001). Eutrophication has been described as the acceleration of the natural aging process of surface waters. It is characterized by excessive loading of silt, organic matter, and nutrients, causing high biological production and decreased basin volume (Cooke et al, 1993). Symptoms of eutrophication (primary impacts) include oxygen supersaturation during the day, oxygen depletion during night, and high sedimentation (filling in) rate. Algae and aquatic plants are the catalysts for these processes. Secondary biological impacts can include loss of biodiversity and structural changes to communities. Phosphorus is generally the nutrient responsible for overfertilization of inland lakes leading to eutrophication.

As reported in the 2002 Integrated List of Waterbodies, the Department identified the following lakes in Northwest Water Region as being eutrophic for a total of 3,480 acres. These four TMDLs will address 2,930 acres or 84.2% of the total impaired acres in this region (Table 2). Lake Hopatcong is listed for both trophic status and aquatic life, which is based on a fishery assessment performed by the Department's Bureau of Freshwater Fisheries; secondary impacts of eutrophication include poorer fish quality and diversity, often due to oxygen depressions and fluctuations. Therefore, it is likely that management actions directed at addressing eutrophication impairments would also address aquatic life impairments based on fishery assessment. However, the exact causes of the aquatic life impairment have not been determined, therefore it is not certain that a TMDL for eutrophication will address the aquatic life impairment completely. Both eutrophic lakes and aquatic life impairments are ranked as Low Priority in the 2002 Integrated List of Waterbodies because they are not directly related to human health issues; however, both issues are environmentally important.

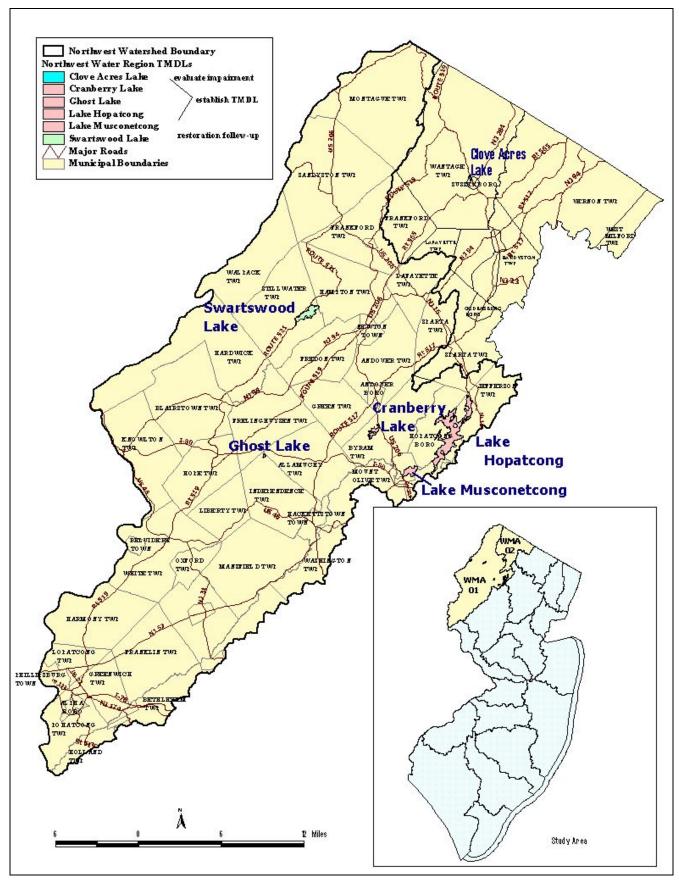
Table 2 Abridged Sublist 5 of the 2002 Integrated List of Waterbodies, eutrophic lakes

| | | | Lake | Lakeshed | |
|-----|-----|-------------------|-------|--------------------|-----------------------|
| No. | WMA | Lake ^a | Acres | Acres | Management Response |
| 1 | 01 | Cranberry Lake | 1890 | 1,740 | establish TMDL |
| 2 | 01 | Ghost Lake | 18.3 | 212 | establish TMDL |
| 3 | 01 | Lake Hopatcong | 2,410 | 16,200 | establish TMDL |
| 4 | 01 | Lake Musconetcong | 314 | 2,980 ^b | establish TMDL |
| 5 | 01 | Swartswood Lake | 521 | 6,410 | restoration follow-up |
| 6 | 02 | Clove Acres Lake | 28.6 | 12,500 | evaluate impairment |

a All of the waterbodies covered under these TMDLs have a FW2 classification.

b To avoid "double-counting," watershed area of Lake Musconetcong does not include Lake Hopatcong and its watershed.

Figure 1 Eutrophic lakes in the Northwest Water Region on Sublist 5 of 2002 Integrated List



These TMDLs will address a total of 2,930 acres of lakes with a corresponding total of 21,110 acres of land. Traditionally, land use has been dictated by the topography and transportation system of the area. The Upper Delaware Watershed (WMA-01) exhibits an accelerated pattern of growth, especially around its lakes. In spite of the area's relatively low population density and numerous protected lands, these development trends are likely to negatively impact surrounding water quality and quantity.

The Department's Geographic Information System (GIS) was used extensively to describe the lakes and lakesheds (watersheds of the lakes), specifically the following data coverages:

- 1995/97 Land use/Land cover Update, published 12/01/2000 by NJDEP Bureau of Geographic Information and Analysis, delineated by watershed management area.
- NJDEP Statewide Lakes (Shapefile) with Name Attributes (from 95/97 Land Use/Land Cover) in New Jersey, published 7/13/2001 by NJDEP Bureau of Freshwater and Biological Monitoring,
 http://www.state.nj.us/dep/gis/digidownload/zips/statewide/njlakes.zip.
- Lakesheds were delineated based on 14-digit hydrologic unit code coverage (HUC-14) elevation contours, and 10 meter digital elevation model grids.
 - NJDEP 14 Digit Hydrologic Unit Code delineations (DEPHUC14), published 4/5/2000 by New Jersey Geological Survey, http://www.state.nj.us/dep/gis/digidownload/zips/statewide/dephuc14.zip.
 - Statewide Elevation Contours (10 Foot Intervals), unpublished, auto-generated from:
 7.5 minute Digital Elevation Models, published 7/1/1979 by U.S. Geological Survey.
 - NJDEP Statewide Elevation Contours (20 Foot Intervals), published 1987 by Bureau of Geographic Information and Analysis (BGIA), http://www.state.nj.us/dep/gis/digidownload/zips/statewide/stcon.zip.
 - NJDEP 10-meter Digital Elevation Grids, published 06/01/2002 by Bureau of Geographic Information and Analysis (BGIA), delineated by watershed management area.
- NJPDES Surface Water Discharges in New Jersey, (1:12,000), published 02/02/2002 by Division of Water Quality (DWQ), Bureau of Point Source Permitting - Region 1 (PSP-R1).

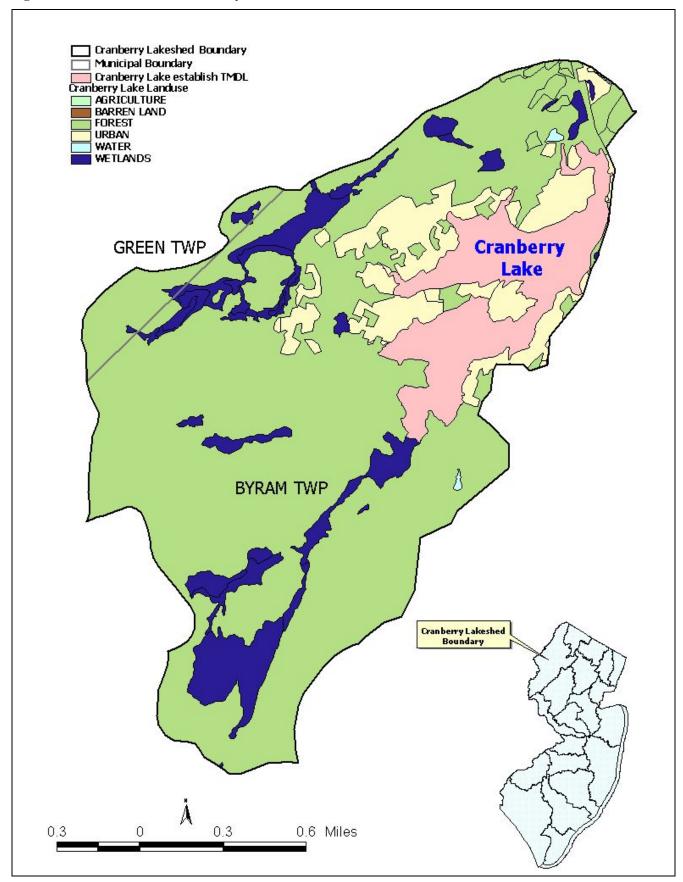
4.1 Cranberry Lake

Cranberry Lake is a 190-acre public lake located in Byram Township, Sussex County, and drains a lakeshed of 1744 acres almost completely within Byram Township. The lakeshed is 9.2 times the area of the lake, making it moderately sized². The lake consists of two basins, each with numerous coves, separated by a large peninsula (Strawberry Point). Tributaries of Lubber's Run feed both basins. Mean depth (2.13m) and total inflow (3,783,000 m³/yr) were obtained from the Diagnostic Feasibility Study for Cranberry Lake (Coastal Environmental Services, 1992).

² A lakeshed seven times the area of its lake is considered small, whereas a lakeshed ten times the area of its lake is considered large (Holdren *et al*, 2001).

Cranberry Lake is a shallow lake within the Musconetcong Watershed, having a mean depth of seven feet and a maximum depth of 15 feet. The lakeshed of Cranberry Lake primarily consists of 1,219 acres of forest, or 70 percent of the entire lakeshed. Approximately half of the land adjacent to the lakeshore is used for medium density residential development, while the remainder is undeveloped (forest cover). Cranberry Lake offers fishing and boating services at the northern end of the lake, where there are boat launch areas (including one trailer launch ramp) and a floating dock. Swimming is available at Cranberry Lake Rose Beach and the Cranberry Lake Club House. The lake is known to have a major septic problem, as indicated in the Phase I Inventory of Current EPA Efforts to Protect Ecosystems report. New Jersey received a Clean Lakes Program Phase II Restoration/Implementation grant in 1992 for Cranberry Lake. The ongoing project entails implementation of in-lake restoration work as well as critical nonpoint source pollution control activities. In 2000 the Weaver House Cove dredging project, which required the lake to be lowered five feet, was nearing completion. In 2001 the Cranberry Lake community was battling watermilfoil vegetation with the herbicide SONAR.

Figure 2 Lakeshed of Cranberry Lake

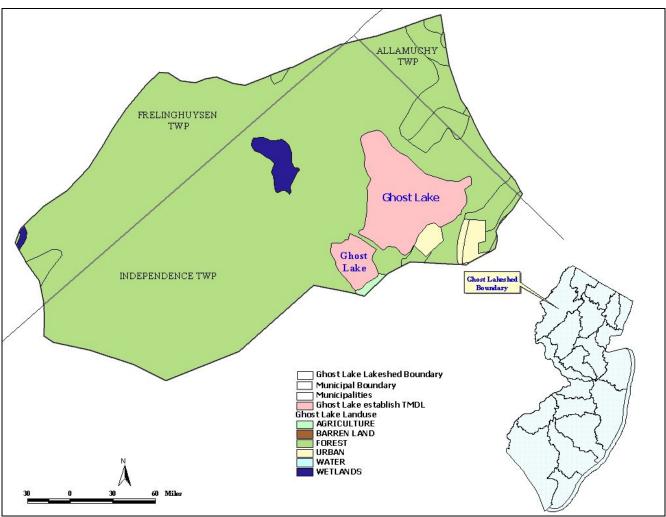


4.2 Ghost Lake

Ghost Lake is located within Jenny Jump State Forest in Independence Township, Warren County, and drains a lakeshed of 212 acres that extends into parts of Frelinghuysen and Allamuchy Townships. The lakeshed is 11.5 times the area of the lakes, making it somewhat large. Ghost Lake has no tributaries; most of the lake's inflow is comprised of groundwater and surface runoff. Mean depth (1.34m) and total inflow (449,000 m³/yr) were obtained from the Phase 1 Diagnostic / Feasibility Study of Ghost Lake (Princeton Hydro, 2002).

This man-made, shallow lake is 18 acres in size and part of the Pequest River Watershed. A narrow band of land separates the larger upper portion of the lake from the much smaller lower portion. The lakeshed is heavily forested and consists of 187 acres of forest (88 percent). However, in the southeast quadrant of the lakeshed, there are two small clusters of low density/rural development, comprising three acres. Ghost Lake offers fishing and boating (car-top launch only) services at the north end of the lake. Although phosphorus loading to Ghost Lake is currently not excessive, a TMDL is being established to ensure that phosphorus levels do not increase in the future and to establish in-lake measures necessary to restore the lake. To that end, the Department plans to implement as appropriate the recommendations in USEPA Clean Lakes Project Phase I Diagnostic/Feasibility Study, dated February 2002.

Figure 3 Lakeshed of Ghost Lake



4.3 Lake Hopatcong

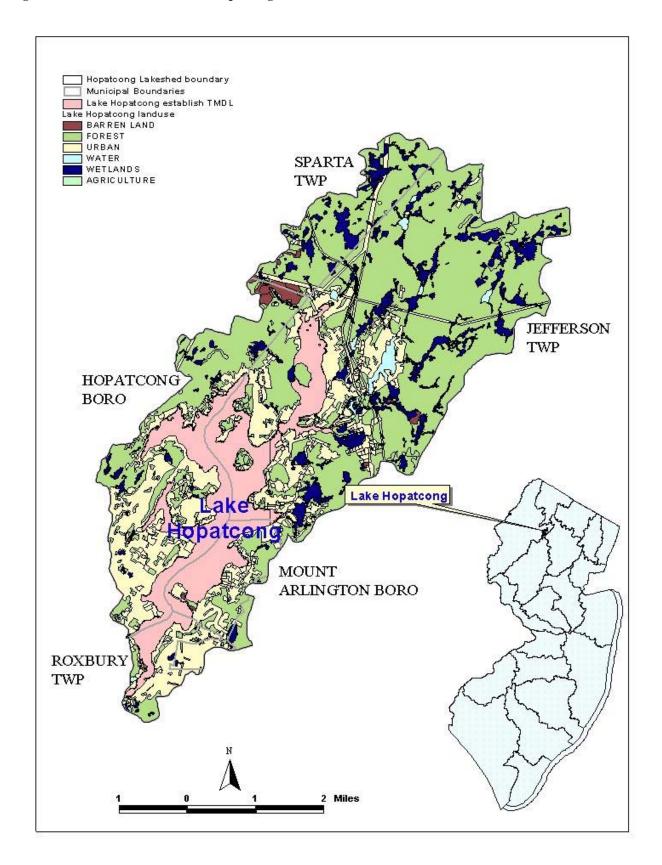
Lake Hopatcong is a 2,406-acre public lake located on the border of Morris and Sussex counties in the municipalities of Mount Arlington Borough, Hopatcong Borough, Jefferson Township, and Roxbury Township. The lake drains a lakeshed of 16,216 acres within the headwaters of the Musconetcong River Watershed. The lakeshed is 6.7 times the area of the lakes, making it fairly small relative to the size of the lake. Lake Hopatcong is a large, irregularly shaped lake composed of many shallow coves around the perimeter. About 50% of the flow into the lake is provided through headwater tributaries of the Musconetcong River, while groudwater inflow comprises about 25% of the flow. Mean depth (5.5m) and total inflow (39,700,000 m³/yr) were obtained from the Clean Lakes Report for Lake Hopatcong (Princeton Aqua Science, 1983).

Lake Hopatcong is the largest freshwater lake in New Jersey and measures 9.5 miles long with a maximum depth of 58 feet. Originally, Hopatcong consisted of two separate lakes, but a dam built in what is now Lake Hopatcong State Park for the Morris Canal Company linked them together in 1837 to form one large lake. Lake Hopatcong was the major source of water

for the 90-mile waterway that stretched from Newark to Phillipsburg. (The lake is also currently a designated emergency source of drinking water.) The predominant land uses in this lakeshed consist of 9,671 acres of forest and wetlands (including bodies of water other than Lake Hopatcong) and 3,974 acres urban, or 60% and 25%, respectively. About 90% of the land adjacent to the 40-mile lake shore is developed, with the majority in seasonal and year-round low, medium, and high density residential land uses. An estimated 500,000 visitors use Lake Hopatcong's recreational facilities each year for fishing, boating, swimming, sailing, jet skiing, and passive recreation. The lake is known to have some good size fish, with largemouth bass averaging two to five pounds. However, several of its beaches are impaired for fecal coliform, and nonpoint source pollution into the lake has become a critical problem.

The Lake Hopatcong Commission is the entity that has taken control of the lake's improvement and was created in January 2001 under a \$3 million startup grant from the State of New Jersey. The Commission's mandate is to safeguard the lake as a natural, scenic, and recreational resource. Some of the tools available to the Commission to accomplish this mandate are monitoring the lake's water quality and quantity; evaluating land use impacts; developing plans, strategies, policies, ordinances, and funding mechanisms; conducting lake management projects; and educating the public on how to protect the lake. A major component of the Lake Hopatcong Commission's protection effort is mechanical harvesting of the overgrowth of aquatic vegetation caused by the influx of phosphorus in the lake. The activity of aquatic plant harvesting began in Lake Hopatcong by the Lake Hopatcong Regional Planning Board and was transferred to the Lake Hopatcong Commission upon its creation. In 2002, the Commission removed over 4.8 million pounds of vegetation from the lake. The Commission has been engaged in a cooperative research effort with the U.S. Geological Survey to conduct a water budget study of Lake Hopatcong. The Commission has initiated plans to address stormwater discharges into Lake Hopatcong, because discharges from such nonpoint sources and catch basins are believed to be a major source of phosphorus in the lake. The Commission also has prepared a geomorphological report describing the Lake Hopatcong area.

Figure 4 Lakeshed of Lake Hopatcong

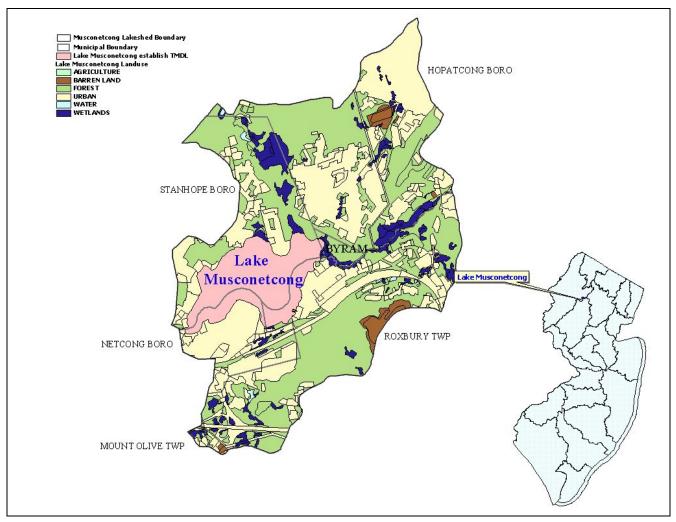


4.4 Lake Musconetcong

Lake Musconetcong is a 314-acre public lake located on the border of Morris and Sussex counties in the municipalities of Netcong, Stanhope, Byram and Roxbury. The lake drains an immediate lakeshed of 2,977 acres within the headwaters of the Musconetcong River Watershed. In addition, Lake Hopatcong drains into the immediate Lake Musconetcong lakeshed, adding 12,091 acres to the total lakeshed. Including the Lake Hopatcong lakeshed, the total lakeshed of Lake Musconetcong is 48 times the area of the lakes, making it very large relative to the size of the lake. Over 80% of the total flow into the lake consists of inflow from Lake Hopatcong through the Musconetcong River. Mean depth (1.5m) and total inflow (48,400,000 m³/yr) were obtained from the Diagnostic / Feasibility Study for Lake Musconetcong (Coastal Environmental Services, 1993).

By far, the predominant land uses in the Lake Musconetcong lakeshed are forest, with 1,199 acres (40%), and urban, with 1,222 acres (41%). Low and medium density residential land uses surround most of the lakeshore itself. Fishing and boating accommodations are offered at the southern end of Lake Musconetcong. There are several municipal and State park areas on the lake available for fishing, one having a boat launch area. Several dozen private docks, but no public ones, can be found on the lake. Stanhope Beach was dredged several years ago, but it is used for fishing only, due to water conditions. The lake is being aided by two entities, the Musconetcong Regional Planning Board, which advises on land use matters affecting the lake, and the Musconetcong Watershed Association. As a result of the Phase I Clean Lakes project, the area immediately around the lake has been sewered and they have implemented weed harvesting, dredging, and various nonpoint source pollution controls. Lake Musconetcong has received Clean Water Act Section 319(h) funding for best management practices.

Figure 5 immediate Lakeshed of Lake Musconetcong



4.5 Swartswood Lake

This glacier lake, with a mean depth of 22 feet (deep lake) and a maximum depth of 42 feet, is 505 acres in size and located at Swartswood State Park in the Paulins Kill Watershed. Fishing and boating services are available on the east and south sides of the lake, and a swimming area is located on the border of the eastern shore and the State Park. The Swartwood Lakes and Watershed Association has been involved in stormwater projects funded by the Clean Water Act Section 319(h) grants to improve water quality, such as the recently completed five-unit hypolimnetic aeration system to halt deterioration and save the trout fishery. The lake had been suffering from large growths of aquatic weeds, algae, and low dissolved oxygen in deeper waters. Now that this public lake is being restored as a result of remediation projects involving weed harvesting, aeration, and nonpoint source controls, the Department agrees to follow up on its restoration to determine whether uses are still impaired. Malcolm Pirnie is currently performing a diagnostic/feasibility study of Swartswood Lake.

4.6 Clove Acres Lake

Clove Acres Lake is located in Wantage Township and Sussex Borough, Sussex County within the Wallkill River Watershed. The initial water quality evaluation for this shallow lake was completed approximately 20 years ago. Subsequently, the dam broke and the 28-acre lake drained. The lake has not been filled for most of the last fifteen years, during which time a limited amount of dredging took place. The dam was rebuilt and, during the summer of 2002, the lake started to refill. The Department will collect information and determine if the new lake is impaired. It should be noted that, based on the previous morphological and hydrological conditions and land use characteristics as of 1995, the Department estimates that the overall nonpoint source load of total phosphorus would have to be reduced by at least 64%. Currently, a consultant Fred Yoerg/Associates has been engaged to begin working on a lake management plan.

5.0 Applicable Surface Water Quality Standards

In order to prevent excessive primary productivity and consequent impairment of recreational, water supply and aquatic life designated uses, the Surface Water Quality Standards (SWQS, N.J.A.C. 7:9B) define both numerical and narrative criteria that address eutrophication in lakes due to overfertilization. The total phosphorous (TP) criterion for freshwater lakes at N.J.A.C. 7:9B – 1.14(c)5 reads as follows:

For freshwater 2 classified lakes, Phosphorus as total phosphorus shall not exceed 0.05 mg/l in any lake, pond or reservoir or in a tributary at the point where it enters such bodies of water, except where site-specific criteria are developed to satisfy N.J.A.C. 7:9B-1.5(g)3.

N.J.A.C. 7:9B-1.5(g)3 states:

The Department may establish site-specific water quality criteria for nutrients in lakes, ponds, reservoirs or stream, in addition to or in place of the criteria in N.J.A.C. 7:9B-1.14, when necessary to protect existing or designated uses. Such criteria shall become part of the SWQS.

Presently, no site-specific criteria apply to any of these lakes.

Also at N.J.A.C. 7:9B-1.5(g)2, the following is discussed:

Except as due to natural conditions, nutrients shall not be allowed in concentrations that cause objectionable algal densities, nuisance aquatic vegetation, or otherwise render the waters unsuitable for the designated uses.

These TMDLs are designed to meet both numeric and narrative criteria of the SWQS.

All of the waterbodies covered under these TMDLs have a FW2 classification. The designated uses, both existing and potential, that have been established by the Department for waters of the State classified as such are as stated below:

In all FW2 waters, the designated uses are (N.J.A.C. 7:9B-1.12):

- 1. Maintenance, migration and propagation of the natural and established aquatic biota;
- 2. Primary and secondary contact recreation;
- 3. Industrial and agricultural water supply;
- 4. Public potable water supply after conventional filtration treatment (a series of processes including filtration, flocculation, coagulation and sedimentation, resulting in substantial particulate removal but no consistent removal of chemical constituents) and disinfection; and
- 5. Any other reasonable uses.

6.0 Source Assessment

Phosphorus sources were characterized on an annual scale (kg TP/yr). Long-term pollutant loads are typically more critical to overall lake water quality than the load at any particular short-term time period (e.g. day). Storage and recycling mechanisms in the lake, such as luxury uptake and sediments dynamics, allow phosphorus to be used as needed regardless of the rate of delivery to the system. Also, empirical lake models use annual loads rather than daily or monthly loads to estimate in-lake concentrations.

6.1 Assessment of Point Sources other than Stormwater

Point sources of phosphorus other than stormwater were identified using the Department's GIS as all Major Municipal (MMJ), Minor Municipal (MMI), and Combined Sewer Overflow (CSO) discharges within each lakeshed. Other types of discharges, such as Industrial, were not included because their contribution, if any, is negligible compared to municipal discharges and runoff from land surfaces. No municipal point sources exist anywhere within the lakesheds of Cranberry Lake or Ghost Lake. One MMI, Arthur Stanlick School, discharges within the Lake Hopatcong lakeshed. The current annual TP load was estimated by multiplying the monthly average TP concentration of 0.314 mg TP/l by the average flow of 0.0014 million gallons per day (MGD), and converting to units of kg/yr. Average flow and concentration were calculated from data submitted to the Department as required in the form of Discharge Monitoring Reports. Similarly, the currently permitted annual TP load was estimated by multiplying the current TP concentration limit of 1.0 mg TP/l by the permitted flow of 0.013 MGD, and converting to units of kg/yr. Since Lake Hopatcong discharges into the Lake Musconetcong lakeshed, the point source was included only indirectly as part of the tributary load into Lake Musconetcong.

Table 3 Point Source Phosphorus Loads

| | | | | current P load | permitted P load |
|----------------|-----------|------------------------|-----------------|-------------------|---------------------|
| Lake | NJPDES# | Facility Name | receiving water | (kg TP/yr) | (kg TP/yr) |
| Lake Hopatcong | NJ0021105 | Arthur Stanlick School | Lake Shawnee | 0.6 | 18.0 |

6.2 Assessment of Nonpoint Sources and Stormwater

Runoff from land surfaces comprises most of the nonpoint and stormwater sources of phosphorus into lakes. Watershed loads for total phosphorus were therefore estimated using the Unit Areal Load (UAL) methodology, which applies pollutant export coefficients obtained from literature sources to the land use patterns within the watershed, as described in USEPA's Clean Lakes Program guidance manual (Reckhow,1979b). Land use was determined using the Department's GIS system using the 1995/1997 land use coverage. The Department reviewed phosphorus export coefficients from an extensive database (Appendix B) and selected the land use categories and values shown in Table 4.

Table 4 Phosphorus export coefficients (Unit Areal Loads)

| land use / land cover | LU/LC codes ³ | UAL (kg TP/ha/yr) |
|-----------------------------------|--------------------------|----------------------|
| medium / high density residential | 1110, 1120, 1150 | 1.6 |
| low density / rural residential | 1130, 1140 | 0.7 |
| Commercial | 1200 | 2.0 |
| Industrial | 1300, 1500 | 1.7 |
| mixed urban / other urban | other urban codes | 1.0 |
| Agricultural | 2000 | 1.5 |
| forest, wetland, water | 4000, 6000, 5000 | 0.1 |
| barren land | 7000 | 0.5 |

Units:

1 hectare (ha) = 2.47 acres

1 kilogram (kg) = 2.2 pounds (lbs)

1 kg/ha/yr = 0.89 lbs/acre/yr

For all lakes in this TMDL document, a UAL of 0.07 kg TP/ha/yr was used to estimate air deposition of phosphorus directly onto the lake surface. This value was developed from statewide mean concentrations of total phosphorus from the New Jersey Air Deposition Network (Eisenreich and Reinfelder, 2001). For Lake Musconetcong, land use runoff loads were only calculated for the immediate watershed downstream of Lake Hopatcong. An additional annual tributary load from Lake Hopatcong into Lake Musconetcong was estimated by multiplying the annual discharge from the lake by the mean phosphorus concentration as calculated under Current Condition in section 7.1 below. Land uses and calculated runoff loading rates for each of the lakes are shown in Table 5. Also included in Table 5 are estimates of loading rates from septic systems and from internal sources (sediment regeneration, macrophyte decomposition, and/or groundwater) developed

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³ LU/LC code is an attribute of the land use coverage that provides the Anderson classification code for the land use. The Anderson classification system is a hierarchical system based on four digits. The four digits represent one to four levels of classification, the first digit being the most general and the fourth digit being the most specific description.

previously (Coastal Environmental Services, 1992; Princeton Aqua Science, 1983; Coastal Environmental Services, 1993; Princeton Hydro, 2002) for each of the lakes.

Table 5 Nonpoint and Stormwater Sources of Phosphorus Loads*

| | | | | | L | ake | | |
|---------------------------------------|--------|----------|---------|----------|-------|----------|-------|----------|
| | Cranbe | rry Lake | Lake Ho | patcong | Musco | netcong | Ghos | st Lake |
| Nonpoint Source | acres | Kg TP/yr | acres | Kg TP/yr | acres | Kg TP/yr | acres | Kg TP/yr |
| medium / high density residential | 156 | 101 | 2,790 | 1,800 | 759 | 492 | 0.0 | 0.0 |
| low density / rural residential | 9.0 | 2.5 | 423 | 120 | 116 | 32.9 | 3.2 | 0.9 |
| commercial | 1.5 | 1.2 | 237 | 192 | 107 | 86.9 | 0.0 | 0.0 |
| industrial | 0.0 | 0.0 | 7.7 | 5.3 | 37.1 | 25.6 | 0.0 | 0.0 |
| mixed urban / other urban | 0.0 | 0.0 | 521 | 211 | 207 | 83.7 | 0.0 | 0.0 |
| agricultural | 0.4 | 0.2 | 0 | 0 | 0.9 | 0.5 | 0.4 | 0.3 |
| forest, wetland, water | 1,380 | 55.9 | 9,670 | 391 | 1,360 | 55.2 | 190 | 7.7 |
| barren land | 6.8 | 1.4 | 165 | 33.3 | 72.6 | 14.7 | 0.0 | 0.0 |
| Direct air deposition on lake surface | 190 | 5.4 | 2,410 | 68.2 | 314 | 8.9 | 18.3 | 0.5 |
| septic systems | n/a | 731 | n/a | 1,600 | r | n/a | 1 | n/a |
| internal load | n/a | 104 | n/a | 595 | n/a | 151 | n/a | 12.4 |
| tributary load | n | /a | n | /a | n/a | 1,240 | ı | n/a |
| TOTAL | 1,740 | 1,000 | 16,200 | 5,020 | 2,980 | 2,190 | 211 | 21.8 |

^{*} all figures rounded to not more than three significant digits

7.0 Water Quality Analysis

Empirical models were used to relate annual phosphorus load and steady-state in-lake concentration of total phosphorus. These empirical models consist of equations derived from simplified mass balances that have been fitted to large datasets of actual lake measurements. The resulting regressions can be applied to lakes that fit within the range of hydrology, morphology and loading of the lakes in the model database. The Department surveyed the commonly used models in Table 6.

Table 6 Empirical models considered by the Department

| reference | steady-state TP concentration in lake (mg/l) | Secondary term | Application |
|-----------------------------------|---|--|---|
| Rast, Jones and Lee, 1983 | $1.81 \times NPL^{0.81}$ | $NPL = \left(\frac{P_a \times DT/D_m}{1 + \sqrt{DT}}\right)$ | expanded database of mostly large lakes |
| Vollenweider and Kerekes, 1982 | $1.22 \times NPL^{0.87}$ | $NPL = \left(\frac{P_a \times \frac{DT}{D_m}}{1 + \sqrt{DT}}\right)$ | mostly large natural lakes |
| Reckhow, 1980 | $\frac{P_a}{13.2}$ | none | Upper bound for closed lake |
| Reckhow, 1979a | $\frac{P_a}{\left(11.6+1.2\times Q_a\right)}$ | $Q_a = \frac{Q_i}{A_l}$ | General north temperate lakes, wide range of loading concentration, areal loading, and water load |

| reference | steady-state TP concentration in lake (mg/l) | Secondary term | Application |
|-----------------------------|---|---|---|
| Walker, 1977 | $\frac{P_a \times DT/D_m}{\left(1 + 0.824 \times DT^{0.454}\right)}$ | none | oxic lakes with $D_m / DT < 50 \mathrm{m/yr}$ |
| Jones and Bachmann, 1976 | $\frac{0.84 \times P_a}{\left(D_m \times \left(0.65 + DT^{-1}\right)\right)}$ | none | may overestimate P in shallow lakes with high $D_{\scriptscriptstyle m}/DT$ |
| Vollenweider, 1975 | $\frac{P_a}{\left(D_m \times \left(DT^{-1} + S\right)\right)}$ | $S = \frac{10}{D_m}$ | Overestimate P lakes with high D_m / DT |
| Dillon-Kirchner, 1975 | $\frac{P_a}{\left(13.2 + \frac{D_m}{DT}\right)}$ | none | low loading concentration range |
| Dillon-Rigler, 1974 | $P_a \times DT/D_m \times (1-R)$ | R = phosphorus retention coefficient | general form |
| Ostrofksy, 1978 | Dillon-Rigler, 1974 | $R = 0.201 \times e^{(-0.0425 \times Q_a)} + 0.5743 \times e^{-0.00949 * Q_a}$ | lakes that flush infrequently |
| Kirchner-Dillon, 1975 | Dillon-Rigler, 1974 | $R = 0.426 \times e^{\left(-0.271 \times D_{m}/DT\right)} + 0.5743 \times e^{-0.00949 \times D_{m}/DT}$ | general application |
| Larsen-Mercier, 1975 | Dillon-Rigler, 1974 | $R = \frac{1}{1 + \sqrt{\frac{1}{DT}}}$ | Unparameterized form |

where: NPL = normalized phosphorus loading

 P_a = areal phosphorus loading $(g/m^2/yr)$

DT = detention time (yr)

 D_m = mean depth (m)

Qa = areal water load $(m/yr)^4$

 Q_i = total inflow (m³/yr)

 A_1 = area of lake (m²)

S = settling rate (per year)

Reckhow (1979a) model was selected because it has the broadest range of hydrologic, morphological and loading characteristics in its database. Also, the model includes an uncertainty estimate that was used to calculate a Margin of Safety. The Reckhow (1979a) model is described in USEPA Clean Lakes guidance documents: Quantitative Techniques for the Assessment of Lake Quality (Reckhow, 1979b) and Modeling Phosphorus Loading and Lake Response Under Uncertainty (Reckhow *et al*, 1980). The derivation of the model is

⁴ Areal water load is defined as the annual water load entering a lake divided by the area of the lake. Since, under steady-state conditions, the water coming in to the lake is equal to the water leaving the lake, either total inflow or total outflow can be used to calculate areal water load. If different values were reported for total inflow and total outflow, the Department used the higher of the two to calculate areal water load.

summarized in Appendix C. The model relates TP load to steady state TP concentration, and is generally applicable to north temperate lakes, which exhibit the following ranges of characteristics (see Symbol definitions after Table 5):

 $phosphorus\ concentration:\ 0.004 < P < 0.135\ mg/l$ average influent phosphorus concentration: $P_a*DT/D_m < 0.298\ mg/l$

areal water load: $0.75 < Q_a < 187 \text{ m/yr}$ areal phosphorus load: $0.07 < P_a < 31.4 \text{ g/m}^2/\text{yr}$

For comparison, Table 7 below summarizes the characteristics for each lake based on their current and target conditions as described below. While the target concentration for each lake (section 7) is well within the range, the areal phosphorus load provides a better representation of a lake's intrinsic loading characteristics. Also, it is the model's prediction of target condition that is being used to calculate the TMDL; if current loads are higher than the range that can produce reliable model results, this has no affect on the model's reliability to predict target condition under reduced loads. It should also be noted that no attempt was made to recalibrate the Reckhow (1979a) model for lakes in New Jersey or in this Water Region, since sufficient lake data were not available to make comparisons with model predictions of steady-state in-lake concentration of total phosphorus. The model was already calibrated to the dataset on which it is based, and is generally applicable to north temperate lakes that exhibit the range of characteristics listed previously.

Table 7 Hydrologic and loading characteristics of lakes

| Lake | Current Avg Influent [TP] (mg/l) | Target Avg Influent [TP] (mg/l) | Current Areal TP load (g/m²/yr) | Target Areal TP load (g/m²/yr) | Areal Water Load (m/year) |
|-------------------|--|---------------------------------------|---------------------------------------|--------------------------------------|------------------------------|
| Cranberry Lake | 0.265 | 0.071 | 1.31 | 0.35 | 4.9 |
| Ghost Lake | 0.049 | 0.049 | 0.19 | 0.19 | 6.0 |
| Lake Hopatcong | 0.126 | 0.080 | 0.52 | 0.33 | 4.1 |
| Lake Musconetcong | 0.045 | 0.030 | 1.73 | 1.14 | 38.1 |

7.1 Current Condition

Using these estimated physical parameters and current loads, the predicted steady-state phosphorus concentration of each lake was calculated using the Reckhow (1979a) formulation and listed in Table 8. The current phosphorus load distribution for each lake is shown in Figures 6 to 9 below.

Figure 6 Current distribution of phosphorus load for Cranberry Lake

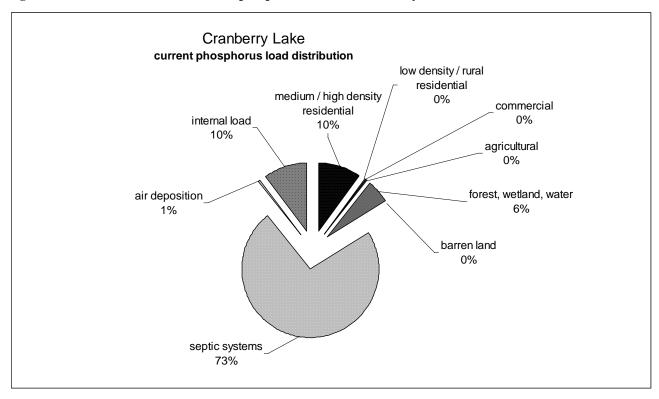


Figure 7 Current distribution of phosphorus load for Ghost Lake

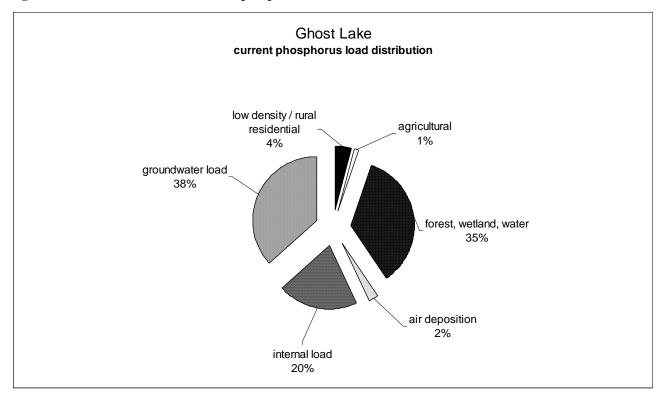


Figure 8 Current distribution of phosphorus load for Lake Hopatcong

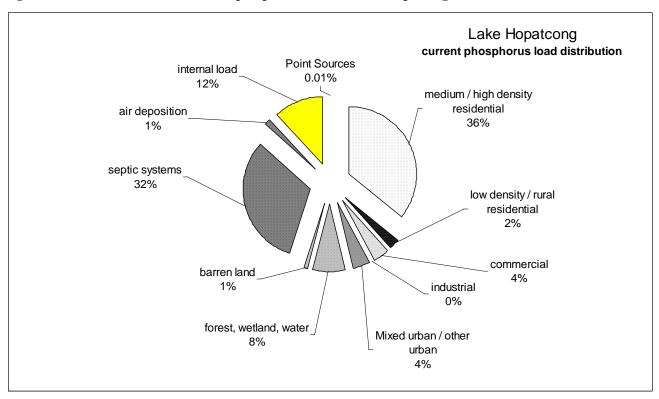
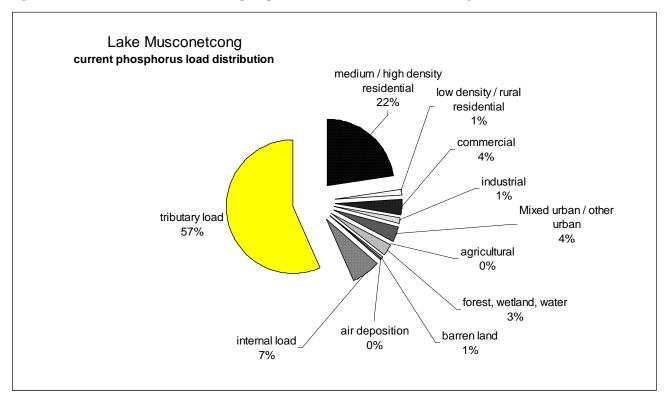


Figure 9 Current distribution of phosphorus load for Lake Musconetcong



7.2 Reference Condition

A reference condition for each lake was estimated by calculating external loads as if the land use throughout the lakeshed were completely forest and wetlands. Using the same physical parameters and external loads from forest and wetlands, a reference steady-state phosphorus concentration was calculated for each lake using the Reckhow (1979a) formulation and listed in Table 8.

7.3 Seasonal Variation/Critical Conditions

These TMDLs will attain applicable surface water quality standards year round. The Reckhow model predicts steady-state phosphorus concentration. To account for data variability, the Department generally interprets threshold criteria as greater than 10% exceedance for the purpose of defining impaired waterbodies. Data from two lakes in New Jersey for which the Department had ready access to data (Strawbridge Lake, NJDEP 2000a; Sylvan Lake, NJDEP 2000b) exhibit peak (based on the 90th percentile) to mean ratios of 1.56 and 1.48, resulting in target phosphorus concentrations of 0.032 and 0.034 mg TP/l, respectively. Since the peak to mean ratios were close and the target concentration not very sensitive to differences in peak to mean ratios, the Department determined that a target phosphorus concentration of 0.03 mg TP/l is reasonably conservative. The seasonal variation was therefore assumed to be 67%, resulting in a target phosphorus concentration of 0.03 mg TP/l. Since it is the annual pollutant load rather than the load at any particular time that determines overall lake water quality (section 6), the target phosphorus concentration of 0.03 mg TP/l accounts for critical conditions.

7.4 Margin of Safety

A Margin of Safety (MOS) is provided to account for "lack of knowledge concerning the relationship between effluent limitations and water quality." (40 CFR 130.7(c)). A MOS is required in order to account for uncertainty in the loading estimates, physical parameters and the model itself. The margin of safety, as described in USEPA guidance (Sutfin, 2002), can be either explicit or implicit (i.e., addressed through conservative assumptions used in establishing the TMDL). For these TMDL calculations, an implicit as well as explicit Margin of Safety (MOS) is provided.

These TMDLs contain an implicit margin of safety by using conservative critical conditions, over-estimated loads, and total phosphorus. Each conservative assumption is further explained below.

Critical conditions are accounted by comparing peak concentrations to mean concentrations and adjusting the target concentration accordingly (0.03 mg TP/l instead of 0.05 mg TP/l). In addition to the conservative approach used for critical conditions, the land use export methodology does not account for the distance between the land use and the lake, which will result in phosphorus reduction due to adsorption onto land surfaces and in-stream kinetic processes. Furthermore, the lakesheds are based on topography without accounting for the

diversion of stormwater from lakes, which is common in urban areas. Neither are any reductions assumed due to the addition of lakeside vegetative buffer construction or other management practices aimed at minimizing phosphorus loads. Finally, the use of total phosphorus, as both the endpoint for the standard and in the loading estimates, is a conservative assumption. Use of total phosphorous does not distinguish readily between dissolved orthophosphorus, which is available for algal growth, and unavailable forms of phosphorus (e.g. particulate). While many forms of phosphorus are converted into orthophosphorus in the lake, many are captured in the sediment, for instance, and never made available for algal uptake.

In addition to the multiple conservative assumptions built in to the calculation, an additional explicit margin of safety was included to account for the uncertainty in the model itself. As described in Reckhow *et al* (1980), the Reckhow (1979a) model has an associated standard error of 0.128, calculated on log-transformed predictions of phosphorus concentrations. Transforming the terms in the model error analysis from Reckhow *et al* (1980) yields the following (Appendix D):

$$MoS_p = \sqrt{\frac{1}{((1-\rho)^*4.5)}} \times (10^{0.128} - 1),$$

where: MoS_p = margin of safety as a percentage over the predicted phosphorus concentration;

 ρ = the probability that the real phosphorus concentration is less than or equal to the predicted phosphorus concentration plus the margin of safety as a concentration.

Setting the probability to 90% yields a margin of safety of 51% when expressed as a percentage over predicted phosphorus concentration or estimated external load. The external load for each lake was therefore multiplied by 1.51 to calculate an "upper bound" estimate of steady-state phosphorus concentration. An additional explicit margin of safety was included in the analyses by setting the upper bound calculations equal to the target phosphorus concentration of 0.3 mg TP/l, as described in the next section and shown in Table 8. Note that the explicit Margin of Safety is equal to 51% when expressed as a percentage over the predicted phosphorus concentration; when expressed as a percentage of total loading capacity, the Margin of Safety is equal to 34%:

$$MoS_{lc} = \frac{MoS_p \times P}{P + (MoS_p \times P)} = \frac{MoS_p}{1 + MoS_p} = \frac{0.51}{1.51} = 0.34$$

where: MoS_p = margin of safety expressed as a percentage over the predicted phosphorus concentration or external load;

 MoS_{lc} = margin of safety as a percentage of total loading capacity; P = predicted phosphorus concentration (or external load).

7.5 Target Condition

As discussed above, the current steady state concentration of phosphorus in each lake must be reduced to a steady state concentration of 0.03 mg/l to avoid exceeding the 0.05 mg/l phosphorus criterion. Using the Reckhow (1979a) formulation, the target conditions were calculated by reducing the loads as necessary to make the upper bound predictions (which incorporate the Margin of Safety) equal to the target phosphorus concentration of 0.03 mg TP/l. The target condition for Ghost Lake was set equal to the current condition, since the upper bound prediction assuming current loads is already less than the target phosphorus concentration of 0.03 mg TP/l. The target condition for Lake Hopatcong was used to calculate the tributary load for the target condition of Lake Musconetcong. Overall reductions necessary to attain the target steady state concentration of total phosphorus in each lake were calculated by comparing the current condition to the target condition (Table 8).

Table 8 Current condition, reference condition, target condition and overall percent reduction for each lake

| Lake | current condition [TP] (mg/l) | reference condition [TP] (mg/l) | upper bound target condition [TP] (mg/l) | target condition [TP] (mg/l) | % overall TP load reduction |
|-------------------|-------------------------------------|---------------------------------------|--|------------------------------------|-----------------------------|
| Cranberry Lake | 0.075 | 0.005 | 0.030 | 0.020 | 73% |
| Ghost Lake | 0.016 | 0.006 | 0.024 | 0.016 | 0% |
| Lake Hopatcong | 0.031 | 0.004 | 0.030 | 0.020 | 36% |
| Lake Musconetcong | 0.030 | 0.011 | 0.030 | 0.020 | 34% |

8.0 TMDL Calculations

8.1 Loading Capacity

The Reckhow (1979a) model was used to solve for loading rate given the upper bound target concentration of 0.03 mg/l (which incorporates the Margin of Safety). Reducing the current loading rates by the percentages in Table 8 yields the same results. The acceptable loading capacity for each lake is provided in Table 10.

8.2 Reserve Capacity

Reserve capacity is an optional means of reserving a portion of the loading capacity to allow for future growth. Reserve capacities are not included at this time. Therefore, the loading capacities and accompanying WLAs and LAs must be attained in consideration of any new sources that may accompany future development. The primary means by which future growth could increase phosphorus load is through the development of forest land within the lakesheds. The implementation plan includes the development of Lake Restoration Plans that require the collection of more detailed information about each lakeshed. If the development of forest with the watershed of a particular lake is planned, the issue of reserve capacity to account for the additional runoff load of phosphorus may be revisited.

8.3 Allocations

USEPA regulations at 40 CFR § 130.2(i), state that "pollutant loadings may be expressed in terms of either mass per time, toxicity, or other appropriate measure." For lake nutrient TMDLs, it is appropriate to express the TMDL on a yearly basis. Long-term average pollutant loadings are typically more critical to overall lake water quality due to the storage and recycling mechanisms in the lake. Also, most available empirical lake models, such as the Reckhow model used in this analysis, use annual loads rather than daily loads to estimate in-lake concentrations.

The TMDLs for total phosphorus are therefore calculated as follows (Table 10):

TMDL = loading capacity

= Sum of the wasteload allocations (WLAs) + load allocations (LAs) + margin of safety + reserve capacity.

WLAs are hereby established for all NIPDES-regulated point sources within each source category, while LAs are established for stormwater sources that are not subject to NJPDES regulation and for all nonpoint sources. This distribution of loading capacity between WLAs and LAs is consistent with recent EPA guidance that clarifies existing regulatory requirements for establishing WLAs for stormwater discharges (Wayland, November 2002). Stormwater discharges are captured within the runoff sources quantified according to land use, as described previously. Distinguishing between regulated and unregulated stormwater is necessary in order to express WLAs and LAs numerically; however, "EPA recognizes that these allocations might be fairly rudimentary because of data limitations and variability within the system." (Wayland, November 2002, p.1) While the Department does not have the data to actually delineate lakesheds according to stormwater drainage areas subject to NIPDES regulation, the land use runoff categories previously defined can be used to estimate between them. Therefore allocations are established according to source categories as shown in Table 9. This demarcation between WLAs and LAs based on land use source categories is not perfect, but it represents the best estimate defined as narrowly as data allow. The Department acknowledges that there may be stormwater sources in the residential, commercial, industrial and mixed urban runoff source categories that are not NJPDESregulated. Nothing in these TMDLs, including Table 9, shall be construed to require the Department to regulate a stormwater source under NJPDES that would not already be regulated as such, nor shall anything in these TMDLs be construed to prevent the Department from regulating a stormwater source under NJPDES. WLAs are hereby established for all NJPDES-regulated point sources, including stormwater, according to their source category. Quantifying WLAs and LAs according to source categories provides the best estimation defined as narrowly as data allow. However it is clearly noted that WLAs are hereby established for all NJPDES-regulated point sources within each source category, while LAs are established for stormwater sources that are not subject to NJPDES regulation and for all nonpoint sources. The WLAs and LAs in Table 9 are not themselves "Additional Measures" under proposed N.J.A.C. 7:14A-25.6 or 25.8.

Table 9 Distribution of WLAs and LAs among source categories

| Source category | TMDL allocation |
|-------------------------------------|-----------------|
| Point Sources other than Stormwater | WLA |
| Nonpoint and Stormwater Sources | |
| medium / high density residential | WLA |
| low density / rural residential | WLA |
| commercial | WLA |
| industrial | WLA |
| Mixed urban / other urban | WLA |
| agricultural | LA |
| forest, wetland, water | LA |
| barren land | LA |
| air deposition onto lake surface | LA |
| septic systems | LA |
| internal load | LA |
| tributary load | LA |

In order to attain the TMDLs, the overall load reductions shown in Table 8, or those determined through additional monitoring, must be achieved. Since loading rates have been defined for at least eight source categories, countless combinations of source reductions could be used to achieve the overall reduction target. The selected scenarios focus on land use and septic sources that can be affected by BMP implementation or NJPDES regulation, requiring equal percent reductions from each in order to achieve the necessary overall load reduction (Table 10). The Lake Restoration Plans developed for each lake as part of the TMDL implementation (section 10) may revisit the distribution of reductions among the various sources in order to better reflect actual implementation projects. The resulting TMDLs, rounded to two significant digits, are shown in Table 10 and illustrated in Figures 10 to 13.

The WLA of 5.5 kg TP/yr for Arthur Stanlick School was calculated by multiplying the current TP concentration limit of 1.0 mg TP/l by the 20-year planned flow of 0.004 MGD, and converting to units of kg/yr. This WLA represents about a 70% decrease from currently permitted annual TP load of 18 kg/yr; however, the actual current annual TP load is only 0.6 kg/yr (section 6.1). Since the WLA represents only 0.1% of the loading capacity for Lake Hopatcong, reduction of the currently permitted concentration limit is not justified. However a WLA was established for this facility in order to prevent the source from becoming significant by incorporating the 20-year planned flow into the next permit. The resulting TMDLs, rounded to two significant digits, are shown in Table 10 and illustrated in Figures 10 to 13.

Table 10 TMDL calculations for each lake (annual loads and percent reductions^a)

| laka | Cranberry Lake | | % | Ghost Lake | | % | |
|-----------------------------------|----------------|---------|-----------|------------|---------|-----------|--|
| lake | kg TP/yr | % of LC | reduction | kg TP/yr | % of LC | reduction | |
| loading capacity (LC) | 400 | 100% | n/a | 33 | 100% | n/a | |
| Point Sources other than | 2/2 | | | | | | |
| Stormwater | n/a | | | n/a | | | |
| Nonpoint and Stormwater | | | | | | | |
| Sources | | | | | | | |
| medium / high density residential | 12 | 3.0% | 88% | 0.00 | 0.0% | n/a | |
| low density / rural residential | 0.30 | 0.08% | 88% | 0.91 | 2.8% | 0% | |
| commercial | 0.15 | 0.04% | 88% | 0.00 | 0.0% | n/a | |
| industrial | 0.00 | 0.00% | n/a | 0.00 | 0.0% | n/a | |
| Mixed urban / other urban | 0.00 | 0.00% | n/a | 0.00 | 0.0% | n/a | |
| agricultural | 0.23 | 0.06% | 0% | 0.27 | 0.81% | 0% | |
| forest, wetland, water | 56 | 14% | 0% | 7.7 | 23% | 0% | |
| barren land | 1.4 | 0.34% | 0% | 0.00 | 0.0% | n/a | |
| air deposition onto lake surface | 5.4 | 1.3% | 0% | 0.52 | 1.6% | 0% | |
| septic systems | 87 | 22% | 88% | | | | |
| internal load | 100 | 26% | 0% | 12 | 38% | 0% | |
| Other Allocations | | | | | | | |
| explicit Margin of Safety | 140 | 34% | n/a | 11 | 34% | n/a | |
| Reserve Capacity | n/a | | | n/a | | | |

| laka | Lake Hopatcong | | % | Lake Musconetcong | | % |
|-----------------------------------|----------------|---------|------------------|-------------------|---------|-----------|
| lake | kg TP/yr | % of LC | reduction | kg TP/yr | % of LC | reduction |
| loading capacity (LC) | 4800 | 100% | n/a | 2200 | 100% | n/a |
| Point Sources other than | 5.5 | 0.11% | 69% ^b | | | |
| Stormwater | | | | | | |
| Nonpoint and Stormwater | | | | | | |
| Sources | | | | | | |
| medium / high density residential | 960 | 20% | 47% | 290 | 13% | 41% |
| low density / rural residential | 64 | 1.3% | 47% | 20 | 0.89% | 41% |
| commercial | 100 | 2.1% | 47% | 52 | 2.4% | 41% |
| industrial | 2.8 | 0.06% | 47% | 15 | 0.69% | 41% |
| Mixed urban / other urban | 110 | 2.3% | 47% | 50 | 2.3% | 41% |
| agricultural | 0.0 | 0.0% | n/a | 0.52 | 0.02% | 0% |
| forest, wetland, water | 390 | 8.1% | 0% | 55 | 2.5% | 0% |
| barren land | 33 | 0.69% | 0% | 15 | 0.67% | 0% |
| air deposition onto lake surface | 68 | 1.4% | 0% | 8.9 | 0.41% | 0% |
| septic systems | 850 | 18% | 47% | | n/a | |
| internal load | 600 | 12% | 0% | 150 | 6.9% | 0% |
| tributary load | | n/a | | 790 | 36% | 36% |
| Other Allocations | | | | | | |
| explicit Margin of Safety | 1600 | 34% | n/a | 740 | 34% | n/a |
| Reserve Capacity | n/a n/a | | | | | |

a Percent reductions shown for individual sources are necessary to achieve overall reductions in Table 7.

b Percent reduction for point source is compared to currently permitted annual load, not actual current load.

Figure 10 Phosphorus allocations for Cranberry Lake TMDL

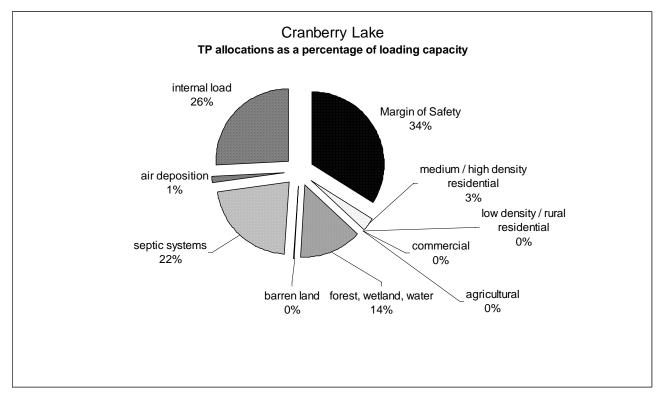


Figure 11 Phosphorus allocations for Ghost Lake

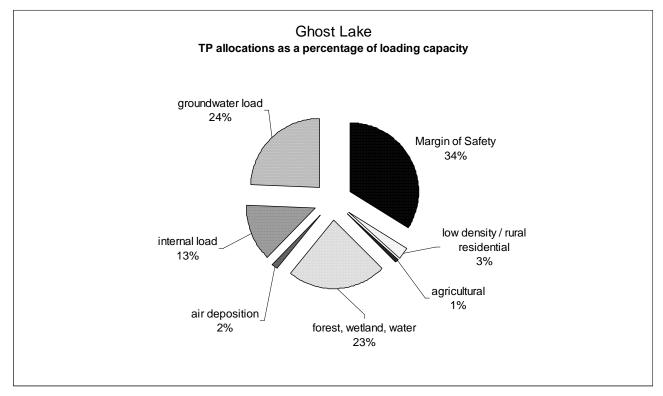


Figure 12 Phosphorus allocations for Lake Hopatcong TMDL

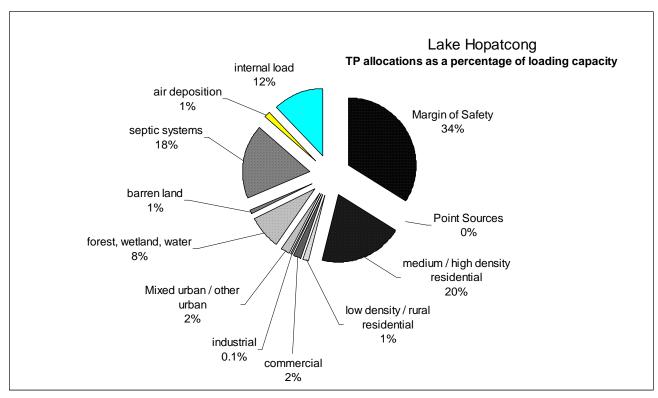
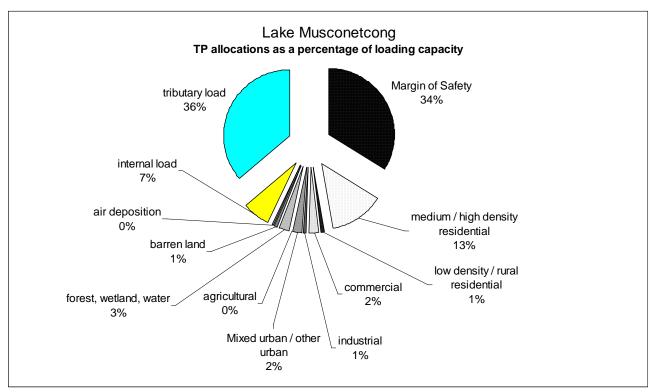


Figure 13 Phosphorus allocations for Lake Musconetcong TMDL



9.0 Follow-up Monitoring

In order to track effectiveness of remediation measures (including TMDLs) and to develop baseline and trend information on lakes, the Department will augment its ambient monitoring program to include lakes on a rotating schedule. The details of a new Lakes Monitoring Network will be published by December 31, 2003. Lakes for which remediation measures have been performed will be given top priority on whatever rotating schedule is developed.

Follow-up monitoring will include evaluations (qualitative using a field index or quantitative) of algal blooms (presence, severity, extent) and aquatic vegetation (density, extent, diversity). Measurements such as secchi depths, nutrient concentrations, and chlorophyll-a will be included, in addition to dissolved oxygen, temperature and pH profiles. Basic hydrologic and morphometric information will be measured as necessary to obtain current data, including discharge and bathymetry. The details as to what data will be collected by the Lakes Monitoring Network will be included in the network description.

10.0 Implementation

The next steps toward implementation are preparation of lake characterizations and lake restoration plans, where they have not already been developed. In the development of these plans, the loads by source will be revised, as necessary, to reflect refinements in source contributions. It will be on the basis of refined source estimates that specific strategies for reduction will be developed. These will consider issues such as cost and feasibility when specifying the reduction target for any source or source type. As appropriate, WLAs or other measures to be applied to traditional or stormwater point sources through NJPDES permits will be adopted by the Department as amendments to the applicable areawide Water Quality Management Plan.

The Department recognizes that TMDLs alone are not sufficient to restore eutrophic lakes. The TMDL establishes the required nutrient reduction targets and provides the regulatory framework to effect those reductions. However, the nutrient load only affects the eutrophication potential of a lake. The implementation plan therefore calls for the collection of additional monitoring data and the development of a Lake Restoration Plan for each lake. The plans will consider in-lake measures that need to be taken to supplement the nutrient reduction measures required by the TMDL. In addition, the plans will consider the ecology of the lake and adjust the eutrophication indicator target as necessary to protect the designated uses.

For instance, with the exception of Lake Hopatcong, all of these lakes are shallow lakes, as defined by having a mean depth less than 3 meters. Even Lake Hopatcong includes many basins that behave like shallow lakes, such as Woodport Bay. For a lake to be shallow means that most of the lake volume is within the photic zone and therefore more able to support aquatic plant growth (Holdren *et al*, 2001). Shallow lakes are generally characterized by either

abundant submerged macrophytes and clear water or by abundant phytoplankton and turbid water. From an aquatic life and biodiversity perspective, it is desirable for shallow lakes to be dominated by aquatic plants rather than algae, especially phytoplankton. While lower nutrient concentrations favor the clear/plant state, either state can persist over a wide range of nutrient concentrations. Shallow lakes have ecological stabilizing mechanisms that tend to resist switches from clear/plant state to turbid/algae state, and vice-versa. The clear/plant state is more stable at lower nutrient concentrations and irreversible at very low nutrient concentrations; the turbid/algae state is more stable at higher nutrient concentrations. The Lake Restoration Plans for each lake will need to consider the ecological nuances of shallow and deep lakes.

The State of New Jersey has adopted a watershed approach to water quality management. That plan divides the state into five watershed management regions, one of which is the Northwest Region. The Department recognizes that lake restoration requires a watershed approach. Lake Restoration Plans will be used as a basis to address overfertilization and sedimentation issues in watersheds that drain to these sensitive lakes. In addition, the Department will direct research funds to understand and demonstrate biomanipulation and other techniques that can be applied in New Jersey lakes to promote the establishment of healthy and diverse aquatic plant communities in shallow lakes. Finally, public education efforts will focus on the benefits of aquatic plants in shallow lakes and the balance of aquatic life uses with recreational uses of these lakes. With the combination of New Jersey's strong commitment to the collection and use of high quality data to support environmental decisions and regulatory programs, including TMDLs, the Department is reasonably assured compliance with the total phosphorus criteria applicable to these eutrophic lakes.

10.1 Lake Characterization

Additional monitoring may be performed in order to develop the Lake Restoration Plans to implement these TMDLs. The level of characterization necessary to plan restoration will be specific to individual lakes depending on the remedial options being considered. During at least one or two summer trips, the following information may be collected as necessary.

- for shallow lakes, vegetation mapping using shore to center transects, measuring density and composition (emergents, rooted floaters, submergents, free-floating plants, submerged macro-algae)
- 1-5 mid-lake sampling stations as needed to characterize the lake
 - o at least 2 samples per station per day; min 4 samples per trip
 - o secchi depths
- chemistry (nutrients, chlorophyll-a, etc.)
 - o surface, metalimnion, hypolimnion, and bottom if stratified
 - o otherwise surface and bottom
- biology (integrated sample from mixed surface layer)
 - o algal abundance and composition (greens, diatoms, blue-greens)
 - o zooplankton abundance, composition and size ranges
- DO, temperature and pH profiles (hourly throughout day)

Where necessary, flow and water quality measurements of influent and effluent streams will be taken periodically from Spring to Fall, and fish abundance and composition will be assessed in early autumn.

The schedules for lake characterization and development of Lake Restoration Plans to implement these TMDLs are provided in Table 11.

Table 11 Implementation Schedule

| Lake | Lake Characterization | Lake Restoration Plan |
|-----------------------------|-------------------------|-------------------------|
| Cranberry Lake ^a | Summer 2004 | Spring 2005 |
| Ghost Lake ^b | Completed 2000 and 2001 | Completed February 2002 |
| Lake Hopatcong | Summer 2003 | Spring 2004 |
| Lake Musconetcong | Summer 2003 | Spring 2004 |

- **a** While Phase 2 remediation of Cranberry Lake is already underway, the TMDL implementation will focus on how successful the nutrient reduction efforts have been, and what additional measures are necessary to restore the lake.
- **b** The Diagnostic / Feasibility study for Ghost Lake (Princeton Hydro, 2002) fulfills the TMDL requirements for lake characterization and lake restoration planning. While nutrient reductions are not required, the report specifies a management plan to restore the lake, including biomanipulation through fishery management.

10.2 Reasonable Assurance

Reasonable assurance for the implementation of these TMDLs has been considered for point and nonpoint sources for which phosphorus load reductions are necessary. These TMDLs obligate the Department to routinely monitor lake water quality as well as characterize and develop specific restoration plan for these particular lakes according to the schedule in Table 11. Moreover, stormwater sources for which WLAs have been established will be regulated as NJPDES point sources.

With the implementation of follow-up monitoring and development of Lake Restoration Plans through watershed management process, the Department is reasonably assured that New Jersey's Surface Water Quality Standards will be attained for these lakes. Activities directed in the watersheds to reduce nutrient loadings shall include a whole host of options, included but not limited to education projects that teach best management practices, approval of projects funded by CWA Section 319 Nonpoint Source (NPS) Grants, recommendations for municipal ordinances regarding feeding of wildlife, and pooper-scooper laws, and stormwater control measures.

11.0 Public Participation

Presently, the Upper Delaware Watershed public participation process is being managed by the Department under a contract with the North Jersey Resource Conservation and Development Council. It is comprised of the Project Work Group, an Action Now Committee, an Education and Outreach Committee, an Open Space and Farmland Preservation Committee, and a Characterization and Assessment Committee. It holds regular meetings and relies on its diverse partners and the general public to work on

watershed tasks and issues. In June 2002 the Department gave a presentation to the Upper Delaware Watershed Project Work Group on the New Jersey 2002 Integrated List of Waterbodies and the Water Quality Monitoring and Assessment Methodology, and also encouraged submittal of any comments.

In accordance with N.J.A.C. 7:15–7.2(g), these TMDLs are hereby proposed by the Department as an amendment to the Upper Delaware Water Quality Management Plan, Northeast Water Quality Management Plan, Upper Raritan Water Quality Management Plan, and Sussex County Water Quality Management Plan. N.J.A.C. 7:15-3.4(g)5 states that when the Department proposes to amend the areawide plan on its own initiative, the Department shall give public notice by publication in a newspaper of general circulation in the planning area, shall send copies of the public notice to the applicable designated planning agency, if any, and may hold a public hearing or request written statements of consent as if the Department were an applicant. The public notice shall also be published in the New Jersey Register.

Notice of these TMDLs was published January 21, 2003 pursuant to the above noted Administrative Code, in order to provide the public an opportunity to review the TMDLs and submit comments. The Department has determined that due to the level of interest in these TMDLs, a public hearing will be held. Public notice of the hearing, provided at least 30 days before the hearing, was published in the New Jersey Register and in two newspapers of general circulation and will be mailed to the applicable designated planning agency, if any, and to each party, if any, who was requested to issue written statement of consents for the amendment.

All comments received during the public notice period and at any public hearings will become part of the record for these TMDLs. All comments will be considered in the establishment of these TMDLs and the ultimate adoption of these TMDLs. When the Department takes final agency action to establish these TMDLs, the final decision and supporting documentation will be sent to U.S.E.P.A. Region 2 for review and approval pursuant to 303(d) of the Clean Water Act (33 U.S.C. 1313(d)) and 40 CFR 130.7.

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Appendix B: Database of Phosphorus Export Coefficients

In December 2001, the Department concluded a contract with the USEPA, Region 2, and a contracting entity, TetraTech, Inc., the purpose of which was to identify export coefficients applicable to New Jersey. As part of that contract, a database of literature values was assembled that includes approximately four-thousand values accompanied by site-specific characteristics such as location, soil type, mean annual rainfall, and site percent-impervious. In conjunction with the database, the contractor reported on recommendations for selecting values for use in New Jersey. Analysis of mean annual rainfall data revealed noticeable trends, and, of the categories analyzed, was shown to have the most influence on the reported export coefficients. Incorporating this and other contractor recommendations, the Department took steps to identify appropriate export values for these TMDLs by first filtering the database to include only those studies whose reported mean annual rainfall was between 40 and 51 inches per year. From the remaining studies, total phosphorus values were selected based on best professional judgement for eight land uses categories.

The sources incorporated in the database include a variety of governmental and non-governmental documents. All values used to develop the database and the total phosphorus values in this document are included in the below reference list.

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Appendix C: Summary of Reckhow (1979a) model derivation

The following general expression for phosphorus mass balance in lake assumes the removal of phosphorus from a lake occurs through two pathways, the outlet (M_0) and the sediments (ϕ) :

$$V \cdot \frac{dP}{dt} = M_i - M_o - \phi$$
 Equation 1

where: $V = lake volume (10^3 m^3)$

P = lake phosphorus concentration (mg/l)
M_i = annual mass influx of phosphorus (kg/yr)
M_i = annual mass offlux of phosphorus (kg/yr)

 M_o = annual mass efflux of phosphorus (kg/yr)

 ϕ = annual net flux of phosphorus to the sediments (kg/yr).

The sediment removal term is a multidimensional variable (dependent on a number of variables) that has been expressed as a phosphorus retention coefficient, a sedimentation coefficient, or an effective settling velocity. All three have been shown to yield similar results; Reckhow's formulation assumes a constant effective settling velocity, which treats sedimentation as an areal sink.

Assuming the lake is completely mixed such that the outflow concentration is the same as the lake concentration, the phosphorus mass balance can be expressed as:

$$V \cdot \frac{dP}{dt} = M_i - v_s \cdot P \cdot A - P \cdot Q$$
 Equation 2

where: $v_s = \text{ effective settling velocity (m/yr)}$

A = area of lake (10^3 m^2)

 $Q = \text{ annual outflow } (10^3 \,\text{m}^3/\text{yr}).$

The steady-state solution of Equation 2 can be expressed as:

$$P = \frac{P_a}{v_s + \frac{Z}{T}} = \frac{P_a}{v_s + Q_a}$$
 Equation 3

where: P_a = areal phosphorus loading rate $(g/m^2/yr)$

z = mean depth (m)

T = hydraulic detention time (yr)

 $Q_a = \frac{Q}{A}$ = areal water load (m/yr).

Using least squares regression on a database of 47 north temperate lakes, Reckhow fit the effective settling velocity using a function of areal water load: $P = \frac{P_a}{11.6 + 1.2 \cdot Q_a}$. Equation 4

Appendix D: Derivation of Margin of Safety from Reckhow et al (1980)

As described in Reckhow *et al* (1980), the Reckhow (1979a) model has an associated standard error of 0.128, calculated on log-transformed predictions of phosphorus concentrations. The model error analysis from Reckhow *et al* (1980) defined the following confidence limits:

$$\begin{split} P_L &= P - h \cdot \left(10^{(\log P - 0.128)} - P \right) \\ P_U &= P + h \cdot \left(10^{(\log P + 0.128)} - P \right) \\ \rho &\ge 1 - \frac{1}{2.25 \cdot h^2} \end{split}$$

where:

 P_L = lower bound phosphorus concentration (mg/l);

 P_U = upper bound phosphorus concentration (mg/l);

P = predicted phosphorus concentration (mg/l);

h = prediction error multiple

 ρ = the probability that the real phosphorus concentration lies within the lower and upper bound phosphorus concentrations, inclusively.

Assuming an even-tailed probability distribution, the probability (ρ_u) that the real phosphorus concentration is less than or equal to the upper bound phosphorus concentration is:

$$\rho_{u} = \rho + \frac{1 - \rho}{2} = \rho + \frac{1}{2} - \frac{\rho}{2} = \rho \cdot \left(1 - \frac{1}{2}\right) + \frac{1}{2} = \frac{1}{2} \cdot \rho + \frac{1}{2}$$

Substituting for ρ as a function of h:

$$\rho_{u} = \frac{1}{2} \cdot \left(1 - \frac{1}{2.25 \cdot h^{2}} \right) + \frac{1}{2} = \frac{1}{2} - \frac{1}{4.5 \cdot h^{2}} + \frac{1}{2} = 1 - \frac{1}{4.5 \cdot h^{2}}$$

Solving for h as a function of the probability that the real phosphorus concentration is less than or equal to the upper bound phosphorus concentration:

$$\frac{1}{4.5 \cdot h^2} = 1 - \rho_u$$

$$h^2 = \frac{1}{4.5(1 - \rho_u)}$$

$$h = \sqrt{\frac{1}{4.5(1 - \rho_u)}}$$

Expressing Margin of Safety (MoS_p) as a percentage over the predicted phosphorus concentration yields:

$$MoS_p = \frac{P_U}{P} - 1 = \frac{P_U - P}{P}$$

Substituting the equation for P_U:

Substituting the equation for
$$P_{U:}$$

$$MoS_{p} = \frac{P + h \cdot (10^{(\log P + 0.128)} - P) - P}{P} = \frac{h \cdot (10^{(\log P + 0.128)} - P)}{P}$$

$$P \cdot MoS_{p} = h \cdot (10^{(\log P + 0.128)} - P)$$

$$\frac{P \cdot MoS_{p}}{h} = 10^{(\log P + 0.128)} - P$$

$$\frac{P \cdot MoS_{p}}{h} + P = 10^{(\log P + 0.128)}$$

Taking the log of both sides and solving for margin of safety:

$$\log\left(\frac{P \cdot MoS_{p}}{h} + P\right) = \log P + 0.128$$

$$\log\left(\frac{P \cdot MoS_{p}}{h} + P\right) - \log P = 0.128$$

$$\log\left(P\left(\frac{MoS_{p}}{h} + 1\right)\right) - \log P = 0.128$$

$$\log P + \log\left(\frac{MoS_{p}}{h} + 1\right) - \log P = 0.128$$

$$\log\left(\frac{MoS_{p}}{h} + 1\right) = 0.128$$

$$\frac{MoS_{p}}{h} + 1 = 10^{0.128}$$

$$\frac{MoS_{p}}{h} = 10^{0.128} - 1$$

$$MoS_{p} = h(10^{0.128} - 1)$$

Finally, substituting for h yields Margin of Safety (MoS_p) as a percentage over the predicted phosphorus concentration, expressed as a function of the probability (ρ_u) that the real phosphorus concentration is less than or equal to the upper bound phosphorus concentration:

$$MoS_p = \sqrt{\frac{1}{((1-\rho_u)^*4.5)}} \times (10^{0.128} - 1)$$