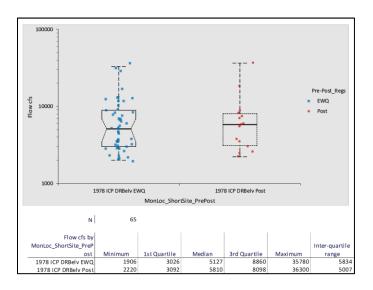
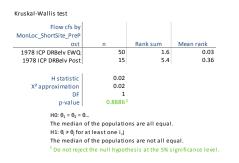


Analysis of flow differences between the EWQ and post-EWQ periods:



Flow was roughly the same between the EWQ and post-EWQ periods. Post-EWQ median flow at Belvidere was about 700 cfs higher than EWQ median flow. Too few samples were collected in the post-EWQ period (n=15). The range of flow conditions sampled was about the same, but more samples would have produced better representation of all flow conditions throughout the range. Considering the under-representation of the flow regime in the post-EWQ data, there is a possibility that water quality differences can falsely interpreted as significant when they really are not. This point is closely considered in each analysis to follow.

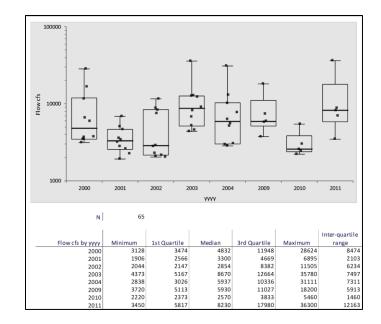


Upstream ICP: Delaware River at Portland 2070 ICP Downstream ICP: Delaware River at Easton 1838 ICP

BCP Watersheds in upstream reach:

Paulins Kill River – 2070 BCP

All other tributaries are less than 20 square miles drainage area.



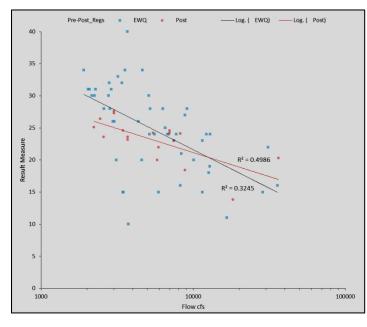
Annual May to September flow statistics associated with water quality measurements are plotted above. These are flow measurements or sometimes estimates associated with the time of each water quality sample. Mean annual flow at this location is about 8,070 cfs; harmonic mean flow is 7460 cfs; and average May to September flow is about 5,770 cfs, which is most typical of summer flow conditions. Though a wide range of flows were sampled by DRBC, these data are most representative of summer flow conditions. Flows corresponding to each water quality sample were taken directly from instantaneous water discharge data from the USGS gage No. 01446500 on the Delaware River at Belvidere.

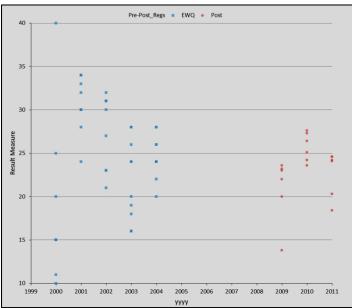
USGS and PADEP maintain a long-term water quality monitoring station at this site. They sample once per quarter every year, whereas DRBC samples twice per month from May through September for three to four year periods. DRBC uses the PADEP/USGS data to check its own results and to supplement the long-term monitoring of PADEP/USGS with more intensive sampling during selected study periods.

Alkalinity as CaCO3, Total mg/l

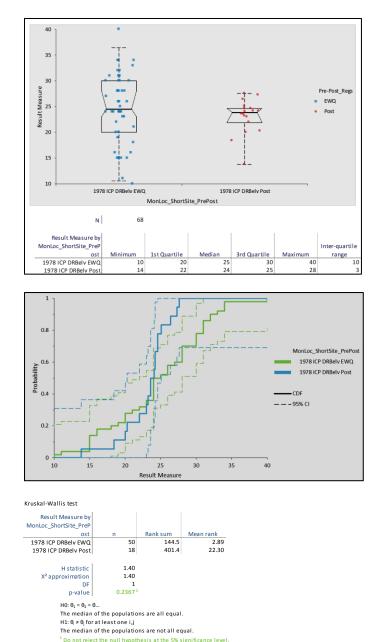
Existing Water Quality (Table 2E):

Median 26 mg/l (recalculated to 25 mg/l, fixed error) Lower 95% Confidence Interval 24 mg/l Upper 95% Confidence Interval 28 mg/l Defined in regulations as a flow-related parameter





No water quality degradation is evident here. Alkalinity apparently did not measurably change between the EWQ and post-EWQ periods.



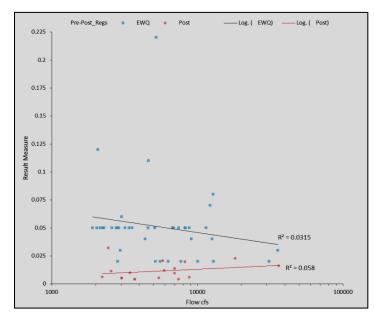
However, sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and insufficient representation of flow conditions. Alkalinity is inversely related to flow in both data sets. Post-EWQ median alkalinity fell within EWQ 95% confidence intervals. Flow is plotted on a logarithmic scale. There were too few samples in the post-EWQ data set, as noted by the pattern of the post-EWQ cumulative distribution function line which not as smooth or gradual as the EWQ line. PADEP and USGS samples were comparable with DRBC results.

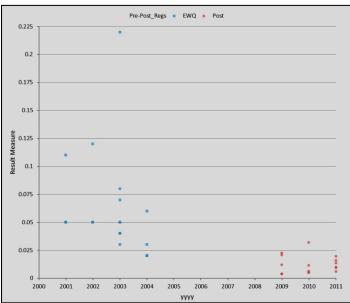
Ammonia Nitrogen as N, Total mg/l

Existing Water Quality (Table 2E):

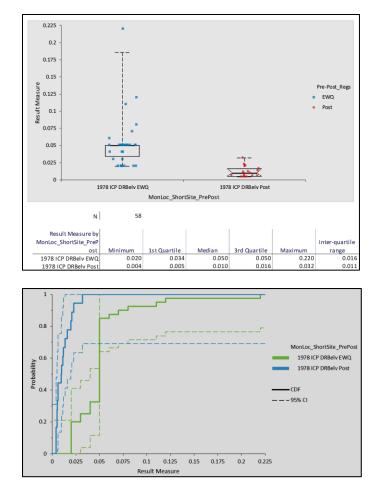
Median <0.05 mg/l

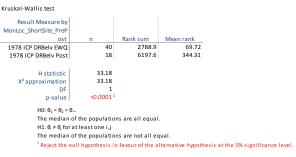
Lower 95% Confidence Interval <0.05 mg/l Upper 95% Confidence Interval <0.05 mg/l





No water quality degradation is evident here. Ammonia concentrations apparently declined. However, sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and insufficient representation of flow conditions.



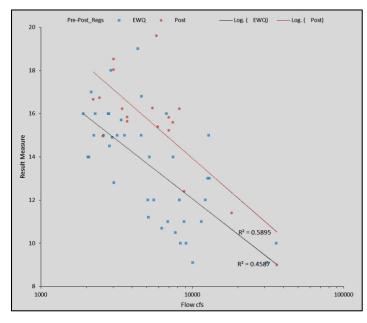


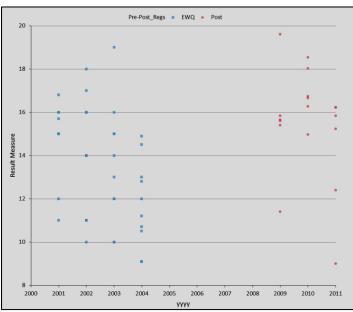
Post-EWQ median ammonia concentration was below the EWQ lower 95% confidence interval. PADEP/USGS data were comparable to DRBC results, with high nondetect frequency and similar concentrations. Flow is plotted on a logarithmic scale. EWQ data included high frequency of undetected results (33 of 40 samples), which interfered with calculation of the median. Under 2009-2011 lower detection levels there were 5/18 undetected results, and the median is a real measurement. Some water quality improvement possibly took place, as the post-EWQ data contained no concentrations greater than 0.032 mg/l, unless the difference is due to laboratory artifacts.

Chloride, Total mg/l

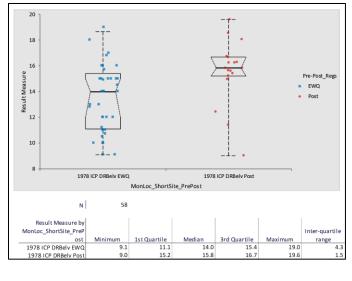
Existing Water Quality (Table 2E):

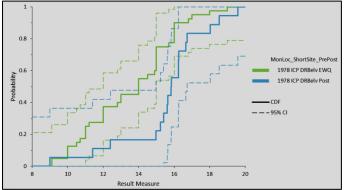
Median 14 mg/l Lower 95% Confidence Interval 12 mg/l Upper 95% Confidence Interval 15 mg/l Defined in regulations as a flow-related parameter

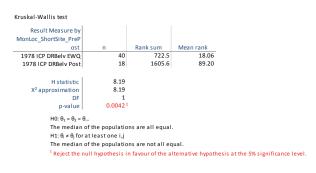




Water quality degradation is evident here. Median chloride concentrations apparently rose by 1.8 mg/l between the two periods.







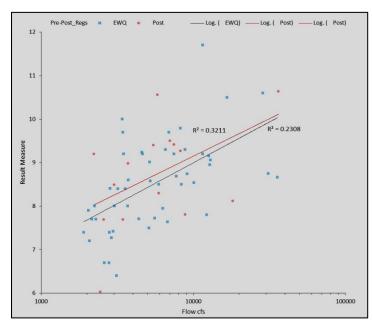
However, sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and insufficient representation of flow conditions. Post-EWQ median concentration rose to just above the EWQ upper 95% confidence interval. There is an obvious separation of about 2 mg/l between the cumulative distributions and concentration vs. flow trends. Chloride concentration is inversely related to flow in both data sets. Flow is plotted on a logarithmic scale. PADEP/USGS data were not numerous enough to validate this conclusion.

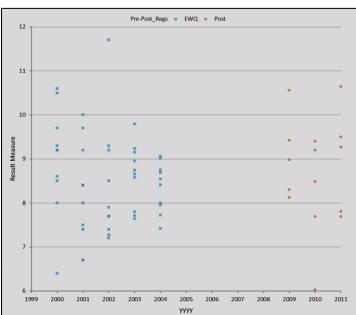
Dissolved Oxygen (DO) mg/l

Existing Water Quality (Table 2E):

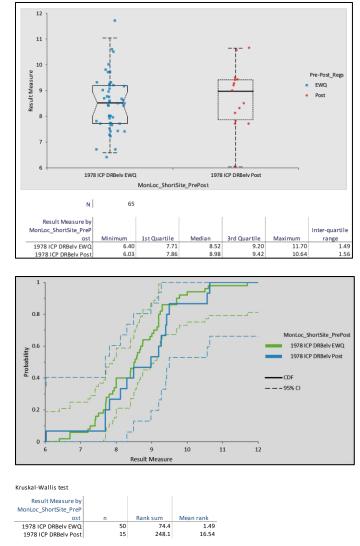
Median 8.52 mg/l

Lower 95% Confidence Interval 8.00 mg/l Upper 95% Confidence Interval 8.95 mg/l





No water quality degradation is evident here. No measurable change took place between the EWQ and Post-EWQ periods.





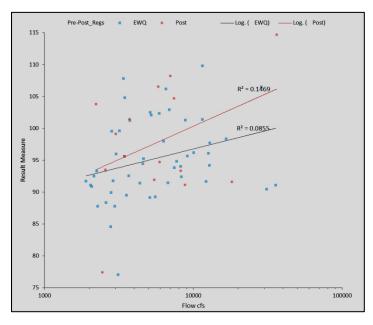
However, analytical uncertainty included insufficient post-EWQ sampling (n=18). Post-EWQ median DO was above the upper EWQ 95% confidence interval, but the increase was not significant and implies water quality improvement. DO is weakly related to flow in both data sets. The site is located on a large pool in the Delaware River, so perhaps oxygen concentrations rise when higher flow conditions produce turbulence that adds oxygen, or it could also be that there are less oxygendemanding pollutants in the pool than there used to be. Flow is plotted on logarithmic scale. The low reading of 6 mg/l in 2010 was probably a probe malfunction. PADEP/USGS data were comparable to DRBC results.

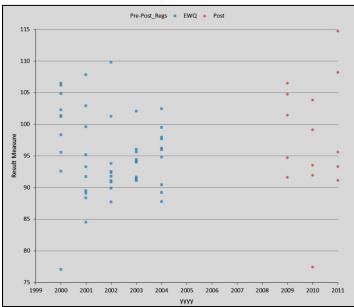
Dissolved Oxygen Saturation %

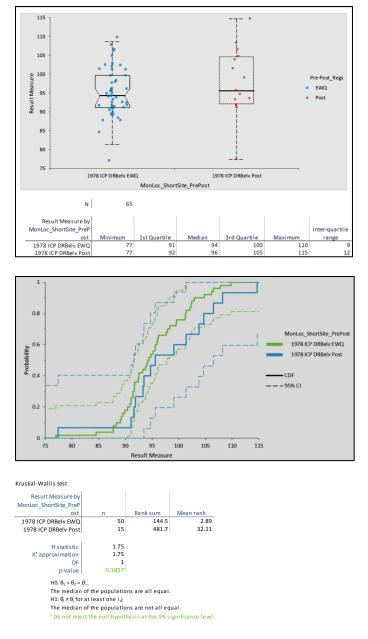
Existing Water Quality (Table 2E):

Median 94%

Lower 95% Confidence Interval 92% Upper 95% Confidence Interval 96%







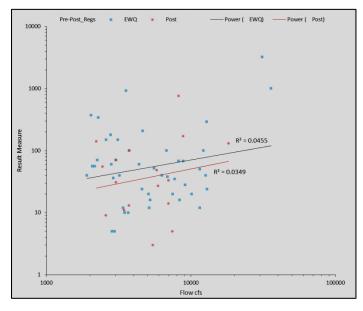
No water quality degradation is evident here. Dissolved Oxygen Saturation is unrelated to flow, and did not measurably change between the EWQ and post-EWQ periods. However, analytical uncertainty included insufficient post-EWQ sampling (n=18). Post-EWQ median DO saturation fell within the EWQ 95% confidence intervals. Flow is plotted on logarithmic scale. There were two low saturation values of 77% and 77.4% found in July 2000 and July 2010, respectively under low flow conditions. Biweekly instead of monthly sampling is recommended for this location. No independent data were available for comparison with DRBC results.

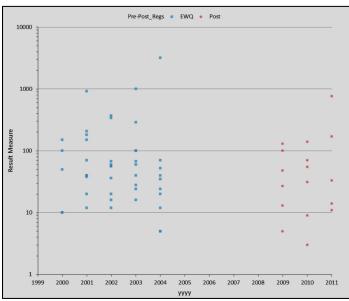
Enterococcus colonies/100 ml

Existing Water Quality (Table 2E):

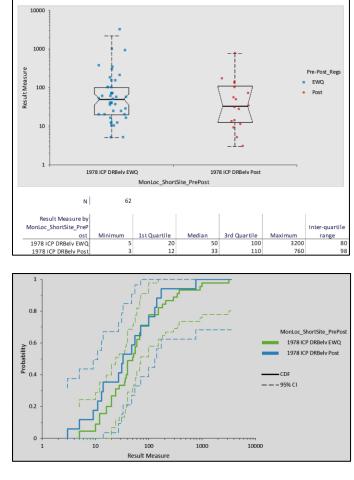
Median 50/100 ml

Lower 95% Confidence Interval 35/100 ml Upper 95% Confidence Interval 68/100 ml





No water quality degradation is evident here. Enterococci apparently declined between the EWQ and Post-EWQ periods, but not significantly.



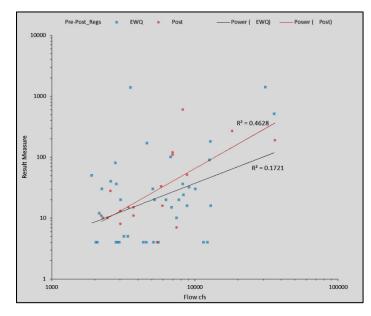


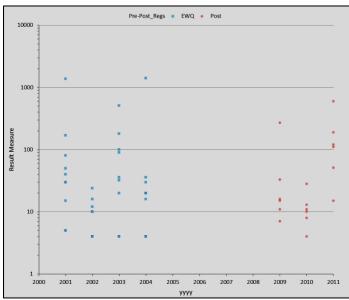
However, sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and insufficient representation of flow conditions. Biweekly instead of monthly sampling is recommended. Enterococcus concentrations are unrelated to flow in both data sets. Concentrations and flows are plotted on logarithmic scale, and regressions are power relationships. Post-EWQ median enterococcus concentrations fell below the lower EWQ 95% confidence interval. No independent data were available for comparison with DRBC results.

Escherichia coli colonies/100 ml

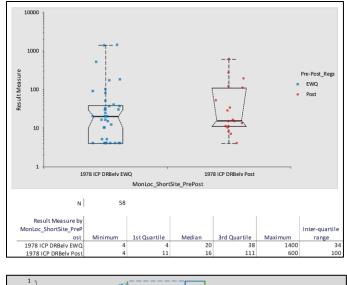
Existing Water Quality (Table 2E):

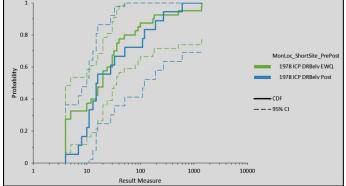
Median 20/100 ml Lower 95% Confidence Interval 5/100 ml Upper 95% Confidence Interval 30/100 ml Defined in rules as a flow-related parameter

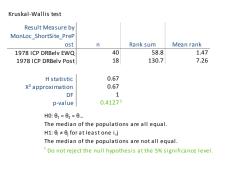




No water quality degradation is evident here. E. coli concentrations apparently did not measurably change between the EWQ and Post-EWQ periods.







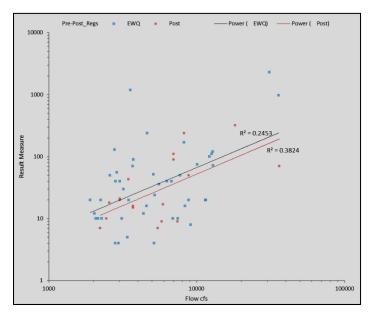
However, sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and insufficient representation of flow conditions. Post-EWQ median E. coli fell within the EWQ 95% confidence intervals. Concentrations and flows are plotted on logarithmic scale, and regressions are power relationships. Biweekly instead of monthly sampling is recommended when assessing this site. E. coli concentrations are weakly related to flow in the EWQ data set, but positively related to flow in the post-EWQ data set. No independent data were available to validate DRBC results.

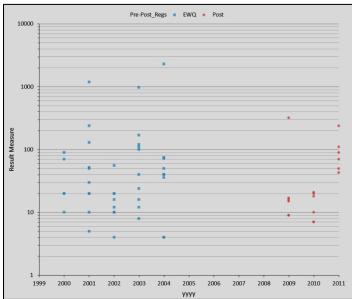
Fecal coliform colonies/100 ml

Existing Water Quality (Table 2E):

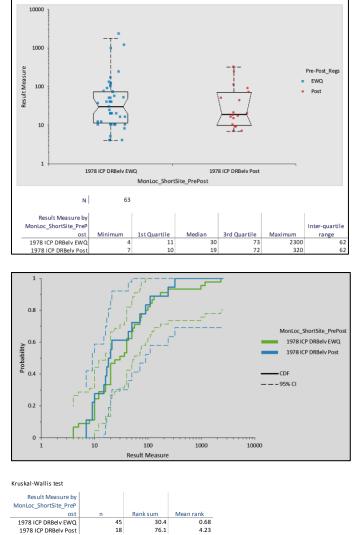
Median 30/100 ml

Lower 95% Confidence Interval 20/100 ml Upper 95% Confidence Interval 50/100 ml





No water quality degradation is evident here. Fecal coliform concentrations fell below the lower EWQ 95% confidence interval, but apparently did not measurably change between the EWQ and post-EWQ periods.



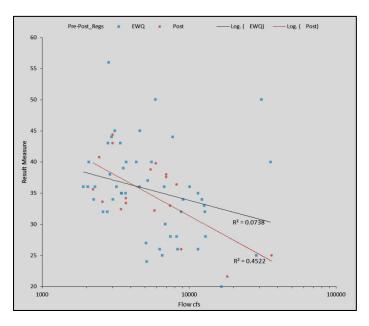


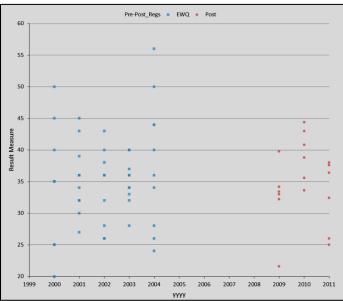
However, sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and insufficient representation of flow conditions. Fecal coliform concentrations are weakly related to flow in both data sets. Biweekly instead of monthly sampling is recommended. Concentrations and flows are plotted on logarithmic scale, and regressions are power relationships. No independent data were available for comparison with DRBC data.

Hardness as CaCO3, Total mg/l

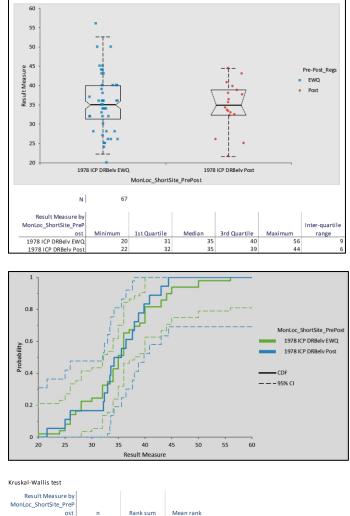
Existing Water Quality (Table 2E):

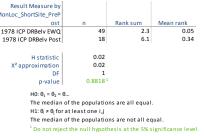
Median 35 mg/l Lower 95% Confidence Interval 33 mg/l Upper 95% Confidence Interval 36 mg/l





No water quality degradation is evident here. Hardness apparently did not measurably change between the EWQ and post-EWQ periods.





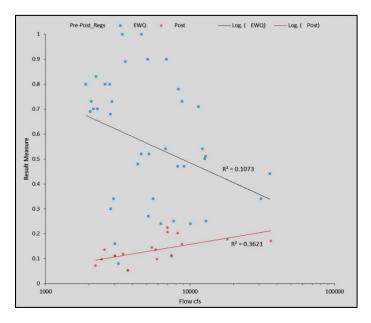
However, sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and insufficient representation of flow conditions. Hardness is unrelated to flow in the EWQ data set, but inversely related to flow in the post-EWQ data set. The strength of the relationship in the post-EWQ data is influenced by only two values, so the relationship is not certain. Post-EWQ median hardness fell within the EWQ 95% confidence intervals, and the cumulative distributions were nearly identical. Flow is plotted on logarithmic scale. USGS/PADEP data were comparable with DRBC results.

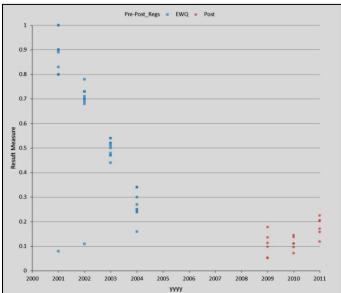
Nitrate + Nitrite as N, Total mg/l

Existing Water Quality (Table 2E, as Nitrate only):

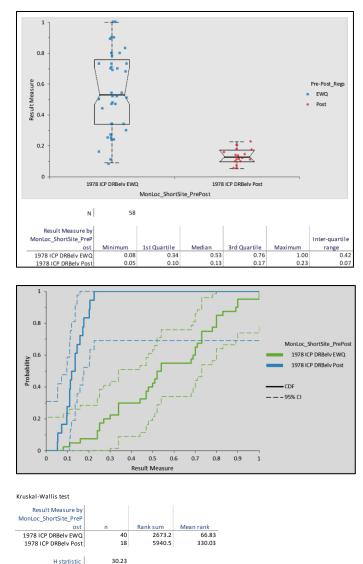
Median 0.53 mg/l

Lower 95% Confidence Interval 0.47 mg/l Upper 95% Confidence Interval 0.71 mg/l





No water quality degradation is evident here. Nitrate concentrations apparently declined between the two periods. Sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and under-representation of flow conditions.





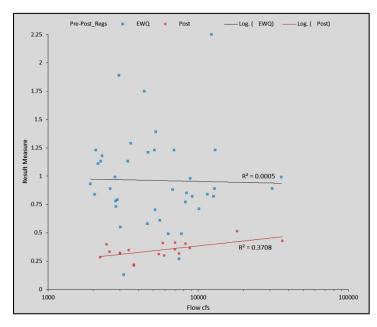
Nitrate is unrelated to flow in EWQ data, but weakly related to flow in post-EWQ data. Post-EWQ nitrate concentrations fell below the lower EWQ 95% confidence interval, and were less variable than EWQ nitrate. Post-EWQ nitrate + nitrite concentrations were assumed equivalent with EWQ nitrate concentrations since EWQ nitrite concentrations were never detected. USGS and PADEP data were most similar to DRBC post-EWQ data, and did not so precipitously decline. However, USGS/PADEP data were also similar to DRBC 2003-2004 data, so perhaps some decline is real and represents an improvement of water quality. Of course the difference could also be mere laboratory artifacts.

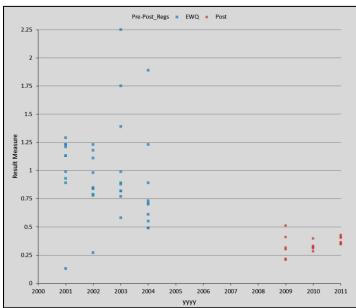
Nitrogen as N, Total (TN) mg/l

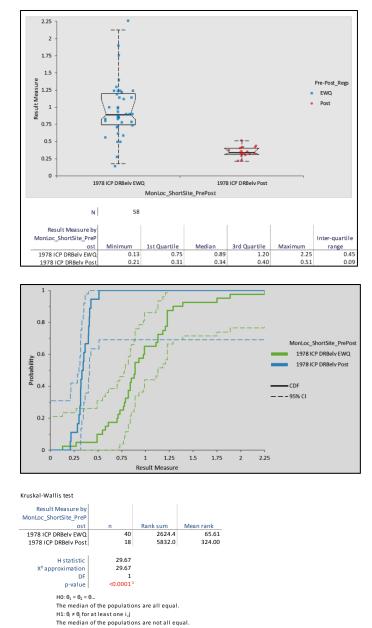
Existing Water Quality (Table 2E):

Median 0.89 mg/l

Lower 95% Confidence Interval 0.82 mg/l Upper 95% Confidence Interval 1.11 mg/l







No water quality degradation is evident here. TN apparently declined between the two periods. Sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and under-representation of flow conditions. TN is unrelated to flow in EWQ data but weakly related in post-EWQ data. Post-EWQ median TN concentrations fell below the EWQ lower 95% confidence intervals, perhaps indicating a water quality improvement. USGS/PADEP data were similar to DRBC 2003-2011 results, showing a slight decline in concentrations but not as drastic as DRBC results indicate.

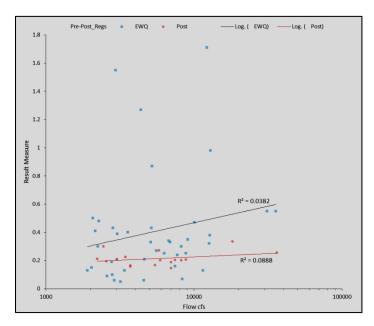
¹ Reject the null hypothesis in favour of the alternative hypothesis at the 5% significance level.

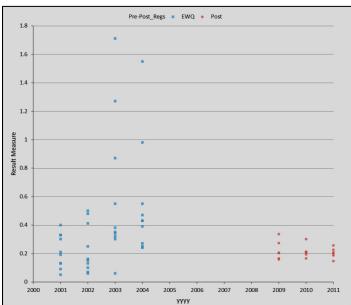
Nitrogen, Kjeldahl as N, Total (TKN) mg/l

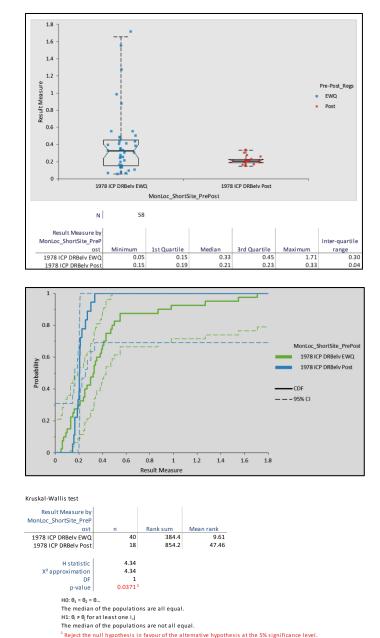
Existing Water Quality (Table 2E):

Median 0.33 mg/l

Lower 95% Confidence Interval 0.24 mg/l Upper 95% Confidence Interval 0.40 mg/l







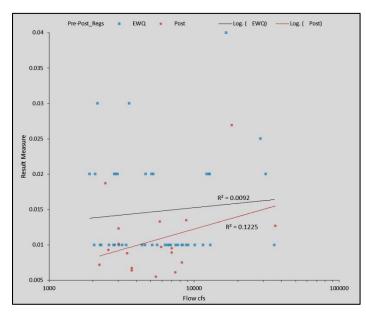
No water quality degradation is evident here. TKN apparently declined between the two periods. Sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and under-representation of flow conditions. The post-EWQ range was far narrower and all concentrations were less than 0.33 mg/l. TKN concentration is unrelated to flow in both data sets. Post-EWQ median TKN fell below the lower EWQ 95% confidence interval. There were no post-EWQ independent data to confirm DRBC results (n=4, all EWQ time frame), but USGS/PADEP values were similar to DRBC results during the EWQ period.

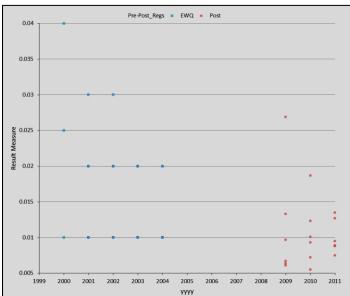
Orthophosphate as P, Total mg/I (OP)

Existing Water Quality (Table 2E):

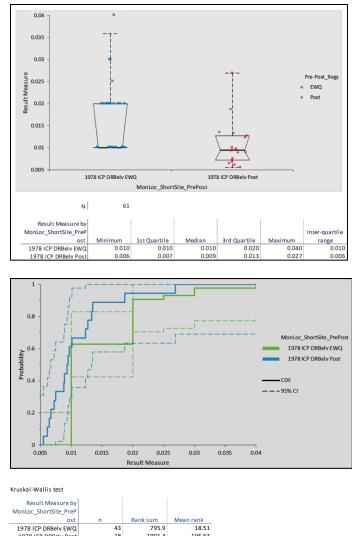
Median <0.01 mg/l

Lower 95% Confidence Interval <0.01 mg/l Upper 95% Confidence Interval 0.02 mg/l





No water quality degradation is evident here. OP apparently declined between the two periods. Sources of analytical uncertainty included potential laboratory artifacts, detection limit differences, insufficient post-EWQ sampling (n=18), and under-representation of flow conditions.



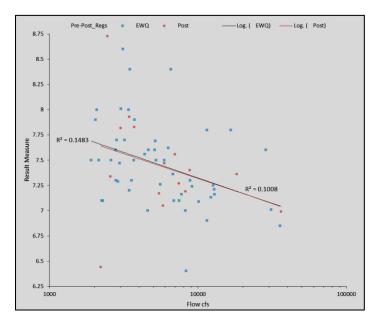


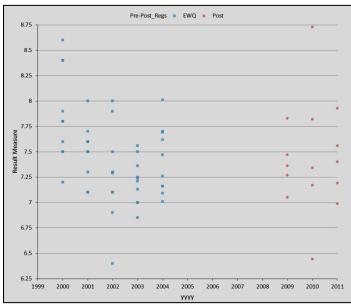
OP is unrelated to flow in both data sets. Post-EWQ median orthophosphate fell below the EWQ lower 95% confidence interval, and the upper quartile of data decreased significantly. This may be due to laboratory artifacts, but perhaps indicates a water quality improvement in that there were no post-EWQ concentrations higher than 0.027 mg/l. The EWQ orthophosphate non-detection rate was 27/43 samples, so the undetected results interfered with estimation of the median and lower confidence interval. Under lower post-EWQ detection limits there were no undetected results. Post-EWQ orthophosphate ranged less widely than EWQ data. USGS/PADEP data were similar to DRBC results. рΗ

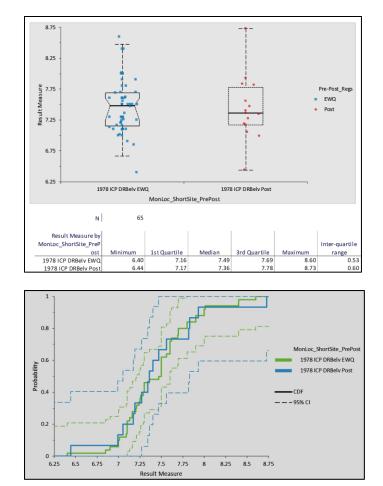
Existing Water Quality (Table 2E):

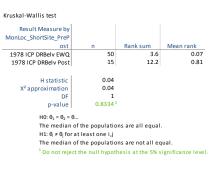
Median 7.49 standard units

Lower 95% Confidence Interval 7.25 standard units Upper 95% Confidence Interval 7.60 standard units









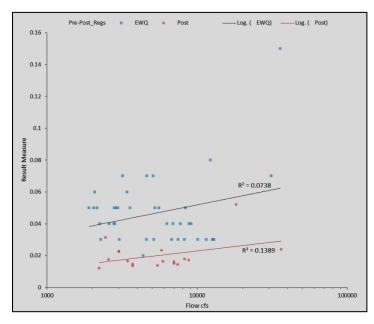
No water quality degradation is evident here. pH apparently did not measurably change between the EWQ and post-EWQ periods. Analytical uncertainty included insufficient post-EWQ sampling (n=18). pH is unrelated to flow in both data sets, but tends toward neutral as flow increases. Post-EWQ median pH was within the EWQ 95% confidence intervals. USGS/PADEP data were similar to DRBC results.

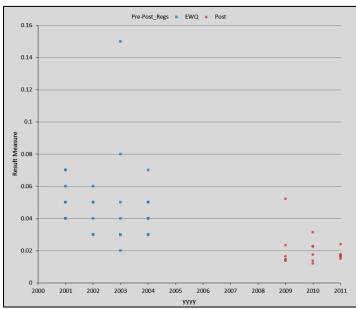
Phosphorus as P, Total (TP) mg/I

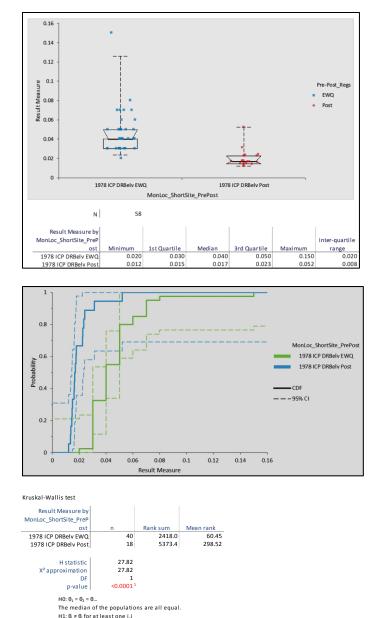
Existing Water Quality (Table 2E):

Median 0.04 mg/l

Lower 95% Confidence Interval 0.04 mg/l Upper 95% Confidence Interval 0.05 mg/l







No water quality degradation is evident here. TP apparently declined between the two periods. Sources of analytical uncertainty included potential laboratory artifacts, detection limit differences, insufficient post-EWQ sampling (n=18), and under-representation of flow conditions. Post-EWQ median total phosphorus fell below the EWQ lower 95% confidence interval. TP is unrelated to flow in both data sets. USGS/PADEP data confirmed DRBC results.

¹ Reject the null hypothesis in favour of the alternative hypothesis at the 5% significance level.

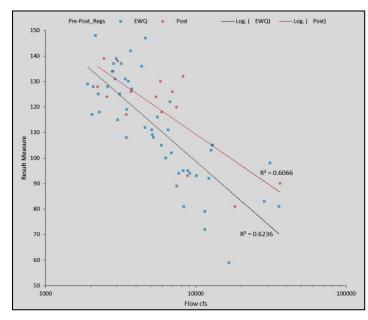
The median of the populations are not all equal.

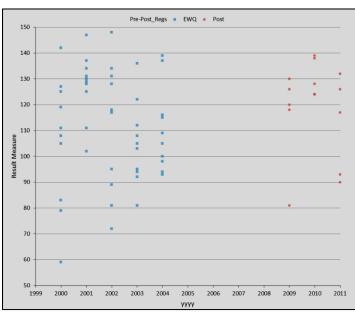
Specific Conductance µmho/cm

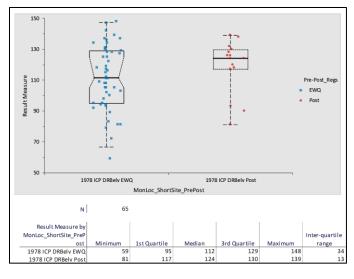
Existing Water Quality (Table 2E):

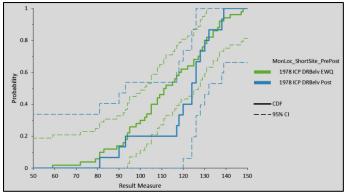
Median 112 μ mho/cm

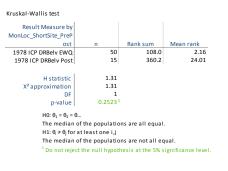
Lower 95% Confidence Interval 105 µmho/cm Upper 95% Confidence Interval 125 µmho/cm Defined in regulations as a flow-related parameter











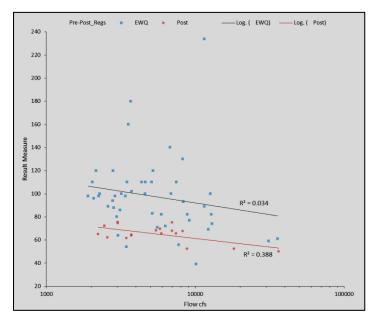
There is slight evidence here of water quality degradation. Specific conductance increased (by about 12 µmho/cm) between the two periods. Analytical uncertainty included insufficient post-EWQ sampling (n=18), and under-representation of flow conditions. Post-EWQ median specific conductance increased near the upper EWQ 95% confidence interval. There were an insufficient number of post-EWQ samples (n=15). Biweekly instead of monthly sampling is recommended here. Specific conductance is inversely related to flow in both data sets. Flow is plotted on logarithmic scale. PADEP data were similar to DRBC results, also showing a slight increase in concentration.

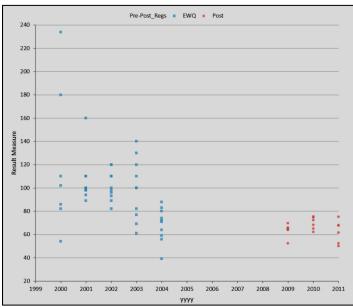
Total Dissolved Solids (TDS) mg/l

Existing Water Quality (Table 2E):

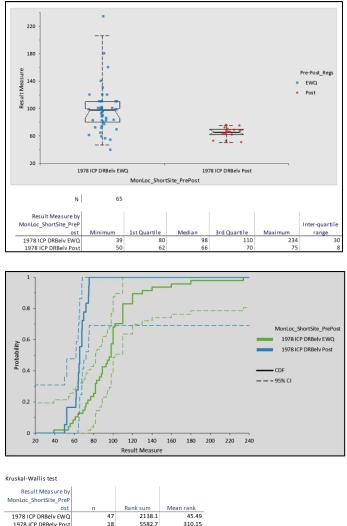
Median 98 mg/l

Lower 95% Confidence Interval 86 mg/l Upper 95% Confidence Interval 100 mg/l





No water quality degradation is evident here. TDS apparently declined between the two periods. Sources of analytical uncertainty included potential laboratory artifacts, insufficient post-EWQ sampling (n=18), and under-representation of flow conditions.



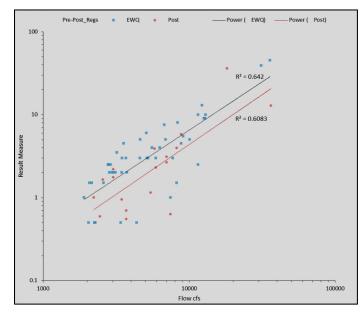


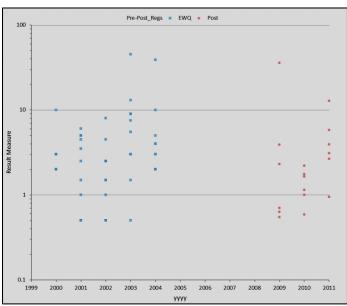
TDS is unrelated to flow in EWQ data but weakly and inversely related in post-EWQ data. Post-EWQ median TDS fell below the EWQ lower 95% lower confidence interval, and was less variable than the baseline samples. Detection limits were different though there were no undetected results at any time. Perhaps this decline is not real but an artifact of different laboratories, though the 2004 data are similar to 2009-2011 data, so some trend may be real. Few PADEP data were available for comparison with results. They were comparable to DRBC results but did not decline.

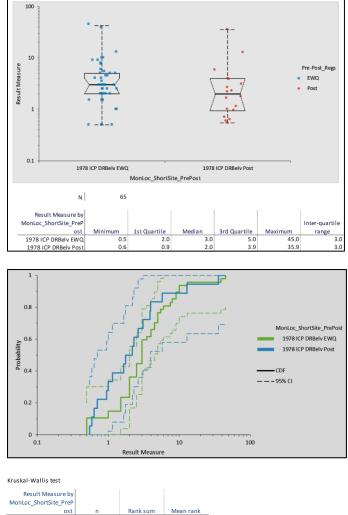
Total Suspended Solids (TSS) mg/l

Existing Water Quality (Table 2E):

Median 3.0 mg/l Lower 95% Confidence Interval 2.0 mg/l Upper 95% Confidence Interval 4.0 mg/l Defined in regulations as a flow-related parameter







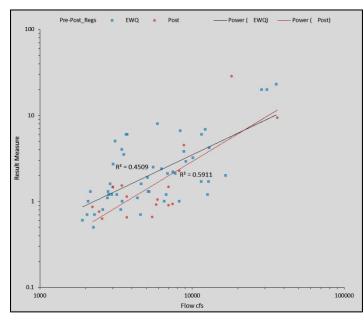


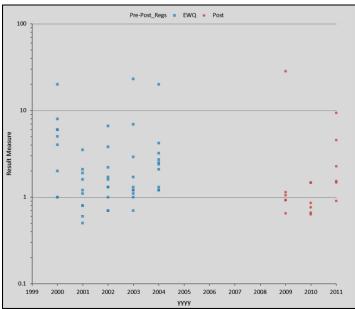
No water quality degradation is evident here. TSS apparently did not measurably change between the EWQ and post-EWQ periods. Sources of analytical uncertainty included potential laboratory artifacts and insufficient post-EWQ sampling (n=18). TSS is positively related to flow in both data sets. Post-EWQ median TSS fell to the lower EWQ 95% confidence interval. Flows and concentrations are plotted on logarithmic scale, and regressions are power relationships. USGS/PADEP data were comparable with DRBC results.

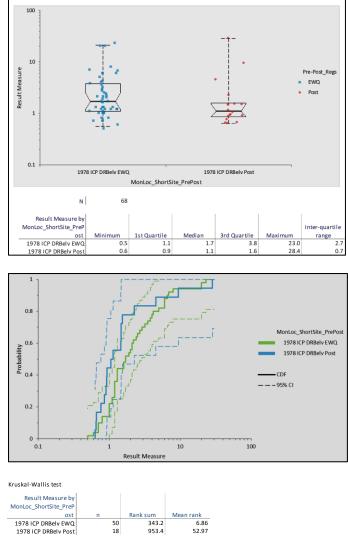
Turbidity NTU

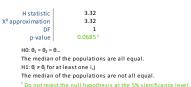
Existing Water Quality (Table 2E):

Median 1.7 NTU Lower 95% Confidence Interval 1.2 NTU Upper 95% Confidence Interval 2.5 NTU Defined in regulations as a flow-related parameter



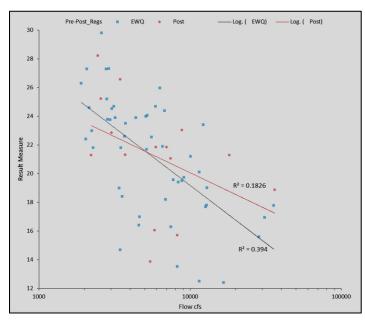


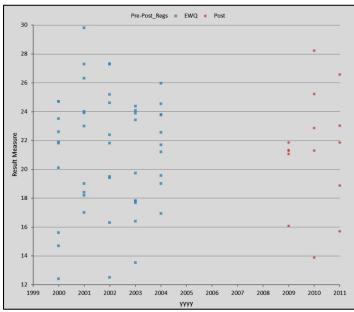


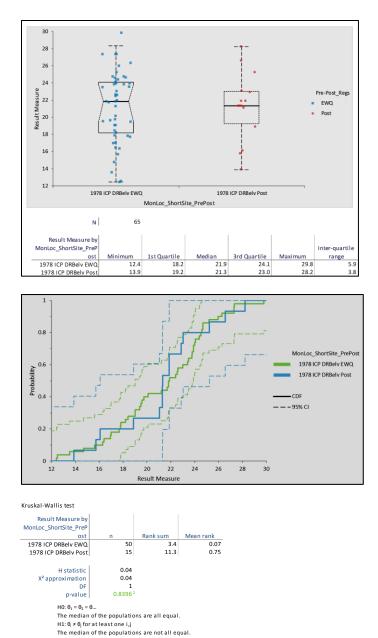


No water quality degradation is evident here. Turbidity apparently declined between the EWQ and post-EWQ periods, but not significantly due to an insufficient number of post-EWQ samples (n=18). The post-EWQ median turbidity fell below the lower EWQ 95% confidence interval. Turbidity is positively related to flow in both data sets; power regression lines are shown. Concentrations and flows are represented on logarithmic scale. There were no independent data available for comparison with DRBC results. Biweekly instead of monthly sampling is recommended at this location. Water Temperature, degrees C

Not included in DRBC Existing Water Quality rules







No water quality degradation is evident here. Water temperature did not measurably change between the EWQ and post-EWQ periods. Analytical uncertainty included insufficient post-EWQ sampling (n=15). Water temperature is inversely related to flow in the post-EWQ data set, but very weakly related to flow in the EWQ data. Flow is plotted on logarithmic scale. PADEP data were comparable with DRBC results.

¹ Do not reject the null hypothesis at the 5% significance level.